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## Metal-Organic-Framework Glasses as Rad Contaminant Sequesters and Nuclear Waste Forms

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**Abstract:** Remediation of Tc remains an unresolved challenge at SRS and other DOE sites. The objective of this project was to develop novel metal organic framework (MOF) glasses for radioactive contaminant sequestration and stabilization from aqueous media. We synthesized, characterized and evaluated ten MOFs of known crystal structure and chemistry for screening MOFs for  $TcO_4^-$  and  $ReO_4^-$  sequestration. MIL-101(Cr) was most effective for  $TcO_4^-$  and  $ReO_4^-$  sequestration from artificial groundwater at equilibration pH values with the removal capacity of 0.076 mg/g for Tc and 0.197 mg/g for Re. MIL-101(Fe), SLUG-22 and UiO-66-NO<sub>2</sub> were moderately effective under similar conditions (pH < 8). These results established a solid foundation for further development of MIL-101(Cr) and new MOFs

to increase  $TcO_4^-$  removal capacity and improve MOF structural stability in FY19. This research may provide a highly applicable platform for solving critical DOE and industrial problems related to nuclear environmental stewardship and nuclear power production.

### FY2018 Objectives

- (1) Synthesize and characterize ten MOFs of known crystal structure and chemistry.
- (2) Evaluate the MOFs for  $TcO_4^-$  and  $ReO_4^-$  removal from artificial groundwater.
- (3) Develop methods for vitrifying the crystalline MOFs.

### Introduction

<sup>99</sup>Tc is a major long-lived fission product created during nuclear power generation. Over the years, Tc has been inadvertently introduced into the environment from leaks at waste storage facilities. <sup>99</sup>Tc currently is one of the key risk drivers at the Savannah River Site (SRS) and other DOE environmental management sites (most notably the Hanford Site, Paducah Gaseous Diffusion Plant, and Oak Ridge National Laboratory). The most common chemical form of Tc in liquid nuclear wastes and in the environment is anionic pertechnetate ( $TcO_4^-$ ).  $TcO_4^-$  displays limited adsorption onto common sediment minerals and is highly mobile making it difficult to capture or to be immobilized [1]. As the stockpile of <sup>99</sup>Tc-bearing nuclear waste continues to increase rapidly, novel sequestration technologies are needed to reduce its potential contamination of the environmental and living organisms.

With current technologies, quaternary amine-based resins have been used to remove aqueous Tc [2]. However, these resins are expensive and have only modest  $TcO_4^-$  loading capacities from the raffinate waste streams. Chemical reductants (e.g., Fe<sub>3</sub>S<sub>4</sub>, soluble or structural Fe(II)) [3] and some bacteria [4] can reduce Tc(VII) to the sparingly soluble Tc(IV). However, the resulting Tc(IV)O<sub>2</sub>·1.6H<sub>2</sub>O has a solubility of 1.5×10<sup>-8</sup> M in groundwater [5], which greatly exceeds the EPA's maximum contaminant level of 5×10<sup>-10</sup> M, and is readily re-oxidized and re-mobilized under most environmental conditions [6]. Tc reduction to form sulfides (e.g., Tc<sub>2</sub>S<sub>7</sub>) [7] or embedding into other sulfide phases [8] or iron oxide waste forms [9] have also been investigated. However, these methods are not practical for many applications. There are

currently no demonstrated technologies that are highly efficient and cost-effective for separation of Tc-containing nuclear waste streams and remediation of aqueous Tc in the contaminated sites.

The overarching objective of this project was to develop novel metal organic framework (MOF) glasses as radioactive contaminant sequestrators and nuclear waste forms. During FY18, we synthesized and characterized ten MOFs of known crystal structure and chemistry, evaluated them for  $\text{TcO}_4^-$  and  $\text{ReO}_4^-$  sequestration from artificial groundwater, and investigated their thermal properties that would help in developing vitrification technology. Based on the selected types of MOFs, we will develop new MOFs that are expected to have higher capacity, improved selectivity and stability for  $\text{TcO}_4^-$  and other anionic contaminant sequestration from aqueous media.

## Approach

A general two-step approach is adopted in the implementation of this project. First, during FY18, we synthesized, characterized and evaluated MOFs of known crystal structure and chemistry to screen out the types of MOFs that might show high capacity and high stability for  $\text{TcO}_4^-$  sequestration from groundwater. As this project has been renewed for FY19, we will focus on selected types of MOFs to develop new MOFs through ion exchange, functionalization, or scissor / insert methods. The new MOFs are expected to have higher Tc capacity and improved stability. Powder X-ray diffraction (XRD), BET surface area measurement and thermogravimetric analysis were used for MOF characterization, while synchrotron radiation X-ray absorption spectroscopy will be applied to studying Tc or Re speciation after the sequestration by MOFs.

## Results/Discussion

### 1. Synthesis of MOFs

We set three criteria in the screening of MOFs for FY18: (i) high stability in aqueous media, including high pH solutions similar to nuclear waste streams, (ii) high capability for removing anions like  $\text{TcO}_4^-$ ,  $\text{I}^-$ ,  $\text{IO}_3^-$ , and organo-I, and (iii) potential for economical vitrification. As such, ten MOFs have been selected and prepared at the University of South Carolina (Dr. Natalia Shustova) to determine which types of MOFs are best fitted to these criteria.

**1.1. ZIF-8** ( $\text{Zn}(\text{mIM})_2$ ; mIM = 2-methylimidazolate, ZIF = Zeolitic Imidazolate Framework). Zinc nitrate hexahydrate ( $\text{Zn}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ ) (0.239 g,  $8.03 \times 10^{-4}$  mol) and 2-methylimidazole (0.060 g,  $7.31 \times 10^{-4}$  mol) was dissolved in 18 mL DMF in a 40 mL vial. The vial was capped and heated at a rate of  $5^\circ\text{C}/\text{min}$  to  $140^\circ\text{C}$  in a programmable oven, held at this temperature for 24 h, then cooled at a rate of  $0.4^\circ\text{C}/\text{min}$  to room temperature. The mother liquor was then decanted and replaced with 20 mL of chloroform. Colorless crystals were collected from the upper layer (floating on the chloroform layer) and washed with DMF (10 mL x 3) and dried under high-vacuum for 1 hour [10].

**1.2. ZIF-67** ( $\text{Co}(\text{mIM})_2$ ; mIM = 2-methylimidazolate). Two separate mixtures of cobalt nitrate hexahydrate ( $\text{Co}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ ) (0.225 g,  $7.73 \times 10^{-4}$  mol) dissolved in 6 mL of deionized  $\text{H}_2\text{O}$ , and 2-methylimidazole (2.75 g,  $3.34 \times 10^{-2}$  mol) dissolved in 40 mL of  $\text{H}_2\text{O}$  were combined in a round-bottom flask equipped with a condenser and heated with stirring at  $60^\circ\text{C}$  for 12 minutes. Stirring was halted after 12 minutes and the solution was left at  $60^\circ\text{C}$  for 20 hours, and then cooled to room temperature. The suspension was then filtered and washed with methanol (10 mL x 3) and dried under high-vacuum for 1 hour [11].

**1.3. ZIF-70** ( $\text{Zn}(\text{IM})_{1.13}(\text{nIM})_{0.87}$ ; IM = imidazolate, nIM = 2-nitroimidazolate). In a 20 mL glass vial, 5 mL from a 2-nitroimidazole stock solution (0.20 M,  $1.0 \times 10^{-3}$  mol) and 5 mL imidazole stock solution (0.20 M,  $1.0 \times 10^{-3}$  mol) were mixed together. Subsequently, 4 mL of a zinc nitrate hexahydrate stock solution (0.20 M,  $8.0 \times 10^{-4}$  mol) was added and stirred. The resulting solution was capped and heated at 110 °C for 4 days, followed by natural cooling to room temperature. Upon cooling, the resulting yellow-orange crystals were filtered and washed with DMF ( $3 \times 15$  mL) [12].

**1.4. UiO-66-NO<sub>2</sub>** ( $\text{Zr}_6\text{O}_4(\text{OH})_4(\text{BDC-NO}_2)_6$ ; BDC-NO<sub>2</sub><sup>2-</sup> = 2-nitrobenzene-1,4-dicarboxylate, UiO = University of Oslo). Zirconium chloride ( $\text{ZrCl}_4$ ) (0.923 g,  $3.96 \times 10^{-3}$  mol) and 2-nitroterephthalic acid (0.845 g,  $4.00 \times 10^{-3}$  mol) were placed in a 40 mL vial and then dissolved in 10 mL of DMF, followed by the addition of 0.67 mL of concentrated hydrochloric acid. The vial was capped and heated at a rate of 1 °C/min to 120 °C in a programmable oven, held at this temperature for 16 h, then cooled at a rate of 0.4 °C/min to room temperature. The powder was then filtered, washed with DMF (10 mL x 3) then placed back in a DMF solution [13].

**1.5. ZIF-71** ( $\text{Zn}(\text{dcIM})_2$ ; dcIM = 4,5-dichloroimidazolate). A solution of zinc acetate dihydrate ( $\text{Zn}(\text{O}_2\text{CCH}_3) \cdot 2\text{H}_2\text{O}$ ) (0.088 g,  $4.0 \times 10^{-4}$  mol) in 15 mL of methanol and a solution of 4,5-dichloroimidazole (0.219 g,  $1.6 \times 10^{-3}$  mol) in 15 mL of methanol were combined in a 40 mL vial. The vial was then sealed and let to stand at room temperature for 24 h. The methanol was then removed by pipet and 20 mL of fresh chloroform was added every day for three days. The crystalline powder was then filtered and dried under high-vacuum for 1 hour [14].

**1.6. ZIF-82** ( $\text{Zn}(\text{nIM})(\text{cIM})$ ; nIM = 2-nitroimidazolate, cIM = 5-cyanoimidazolate). In a 20 mL glass vial, 5 mL of a 2-nitroimidazole stock solution (0.20 M,  $1.0 \times 10^{-3}$  mol) and 5 mL of a 5-cyanoimidazole stock solution (0.20 M,  $1.0 \times 10^{-3}$  mol) were mixed together. Subsequently, 4 mL of a zinc nitrate hexahydrate stock solution (0.20 M,  $8.0 \times 10^{-4}$  mol) was added and stirred. The resulting solution was capped and heated at 100 °C for 4 days, followed by natural cooling to room temperature. Upon cooling, the resulting yellow-orange crystals were filtered and washed with DMF ( $3 \times 15$  mL) [12].

**1.7. MIL-101-Cr** ( $\text{Cr}_3\text{F}(\text{H}_2\text{O})_2\text{O}(\text{BDC})_3$ ; BDC = 1,4-benzenedicarboxylate, MIL = Materials Institute Lavoisier). Chromium(III) nitrate nonahydrate (0.400 g,  $1.00 \times 10^{-3}$  mol) was dissolved in 4.8 mL of deionized water, then placed in a Teflon-lined autoclave. Terephthalic acid (0.164 g,  $9.87 \times 10^{-4}$  mol) was then added followed by the addition of 0.2 mL of HF (5.0 M,  $1.0 \times 10^{-3}$  mol). The autoclave was then sealed and placed in an oven at 220 °C for 8 hours. Upon cooling, the resulting green solid was centrifuged to collect the precipitate and solvent exchanged with DMF ( $3 \times 10$  mL) with centrifugation between each exchange. The product was then heated in DMF (120 °C, 12 h), filtered, and then boiling ethanol (80 °C, 12 h) and filtered off [15].

**1.8. MIL-101-Fe** ( $\text{Fe}_3\text{Cl}(\text{C}_3\text{H}_7\text{NO})\text{O}(\text{BDC})_3$ ; BDC = 1,4-benzenedicarboxylate). Iron(III) chloride hexahydrate (0.325 g,  $1.20 \times 10^{-3}$  mol) and terephthalic acid (0.103 g,  $6.20 \times 10^{-4}$  mol) were dissolved in 10 mL DMF by sonication, followed by stirring for 20 minutes. The resulting solution was then loaded into a Teflon-lined autoclave. The autoclave was then sealed and placed in an oven at 110 °C for 24 hours. Upon cooling, the resulting brown solid was centrifuged to collect the precipitate and solvent exchanged with DMF ( $3 \times 10$  mL) with centrifugation between each exchange. The product was then heated in DMF (120 °C, 12 h), filtered, and then boiling ethanol (80 °C, 12 h) and filtered off [15].

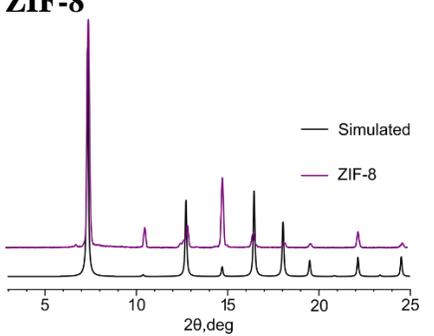
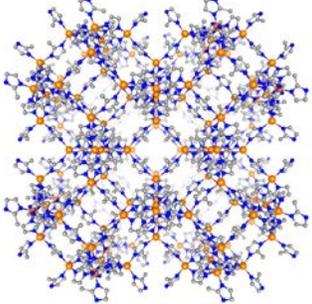
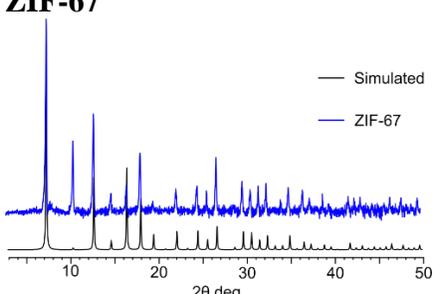
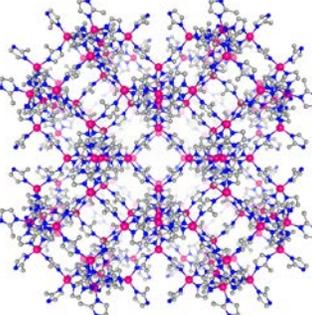
**1.9. SLUG-21** ( $\text{Ag}_2(\text{bpy})_2(\text{O}_3\text{SCH}_3\text{CH}_2\text{SO}_3)$ ; bpy = 4,4-bipyridine, SLUG = University of California, Santa Cruz). Silver nitrate ( $\text{AgNO}_3$ , 0.236 g,  $1.40 \times 10^{-3}$  mol), 1,3-ethanedithionic acid dehydrate (0.314 g,  $1.39 \times 10^{-3}$  mol) and 4,4-bipyridine (0.217 g,  $1.39 \times 10^{-3}$  mol) were

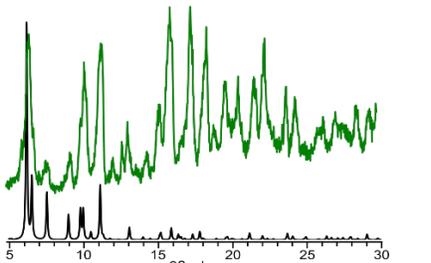
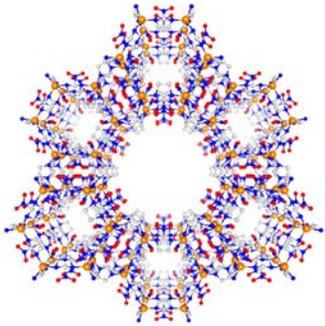
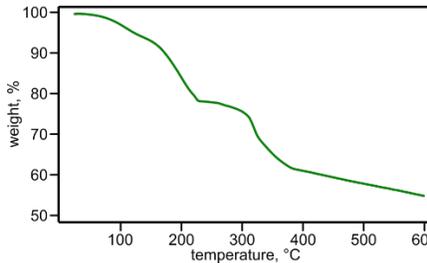
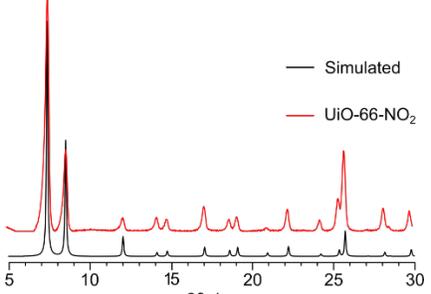
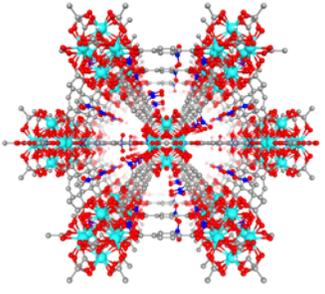
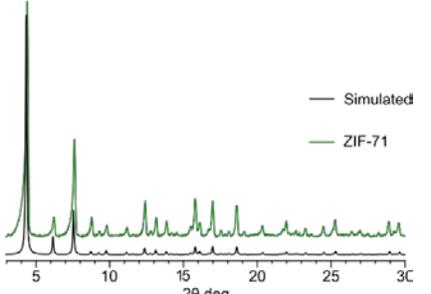
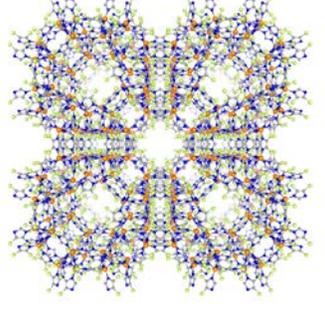
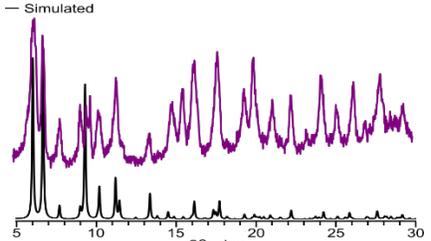
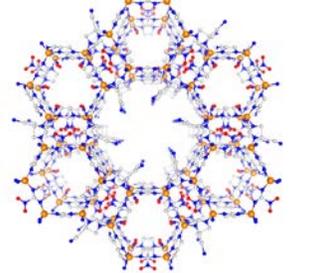
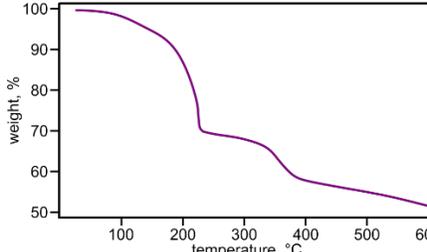
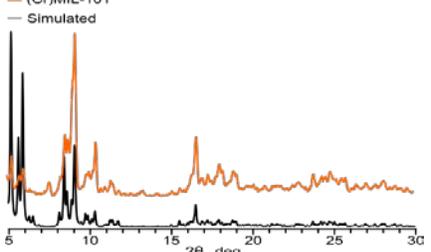
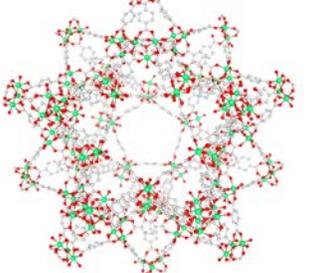
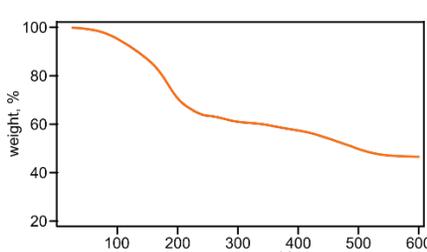
suspended in 10 mL of deionized water and stirred for 3 minutes. The resulting mixture was loaded into a Teflon-lined autoclave, sealed, and then heated in an oven at 150 °C for 3 days, followed by natural cooling to room temperature. Upon cooling, the resulting colorless plate crystals were filtered, washed with deionized water ( $3 \times 15$  mL) and washed with acetone ( $3 \times 15$  mL) [16].

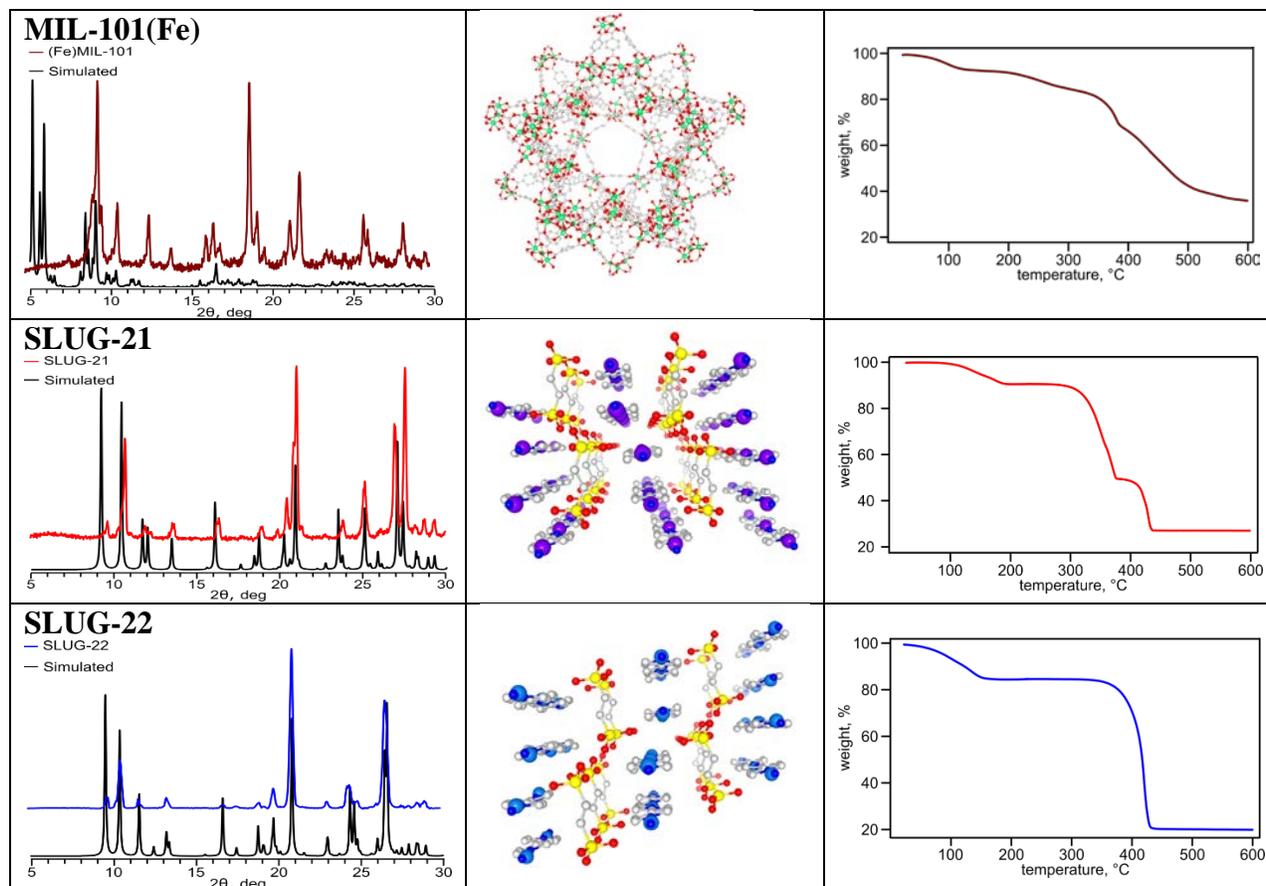
**1.10. SLUG-22** ( $\text{Cu}_2(\text{bpy})_2(\text{O}_3\text{SCH}_3\text{CH}_2\text{SO}_3)$ ; bpy = 4,4-bipyridine). Copper acetate monohydrate (0.054 g,  $2.70 \times 10^{-4}$  mol), 1,2-ethanedisulfonic acid dihydrate (0.073 g,  $3.23 \times 10^{-4}$  mol) and 4,4'-bipyridine (0.042 g,  $2.70 \times 10^{-4}$  mol) were suspended in 2 mL of deionized water and stirred for 30 minutes. The resulting mixture was loaded into a Teflon-lined autoclave, sealed, and then heated in an oven at 175 °C for 4 days, followed by cooling at a rate of 7 °C/hour. Upon cooling, the resulting yellow-green needle-like crystals were filtered, washed with deionized water ( $3 \times 15$  mL) and washed with acetone ( $3 \times 15$  mL) [17].

## 2. Characterization of MOFs

The MOFs were characterized by powder X-ray diffraction (XRD) and thermogravimetric analysis (TGA) to confirm their known crystal structure (Figure 1).

A. XRD	B. Structure	C. TGA
<p><b>ZIF-8</b></p> 		N/A
<p><b>ZIF-67</b></p> 		N/A

<p><b>ZIF-70</b></p> <p>— ZIF-70 — Simulated</p> 		
<p><b>UiO-66-NO<sub>2</sub></b></p> <p>— Simulated — UiO-66-NO<sub>2</sub></p> 		<p>N/A</p>
<p><b>ZIF-71</b></p> <p>— Simulated — ZIF-71</p> 		<p>N/A</p>
<p><b>ZIF-82</b></p> <p>— ZIF-82 — Simulated</p> 		
<p><b>MIL-101(Cr)</b></p> <p>— (Cr)MIL-101 — Simulated</p> 		



**Figure 1. XRD (A), crystal structure (B) and TGA (C) of MOFs**

### 3. Effectiveness of MOFs for $\text{TcO}_4^-$ and $\text{ReO}_4^-$ sequestration from artificial groundwater

The ten MOFs were evaluated for  $\text{TcO}_4^-$  and its non-radioactive surrogate,  $\text{ReO}_4^-$ , sequestration from artificial groundwater under atmospheric ( $P_{\text{CO}_2} = 10^{-3.5}$  atm) conditions. Sorption coefficient ( $K_d$ , mL/g) and the equilibrium sorption capacity ( $q_e$ , mg/g) were calculated using formula 1 and 2, respectively:

$$K_d = \frac{C_0 - C_e}{C_e} \times \frac{V}{M} \quad (1)$$

$$q_e = \frac{(C_0 - C_e) \times V}{M} \quad (2)$$

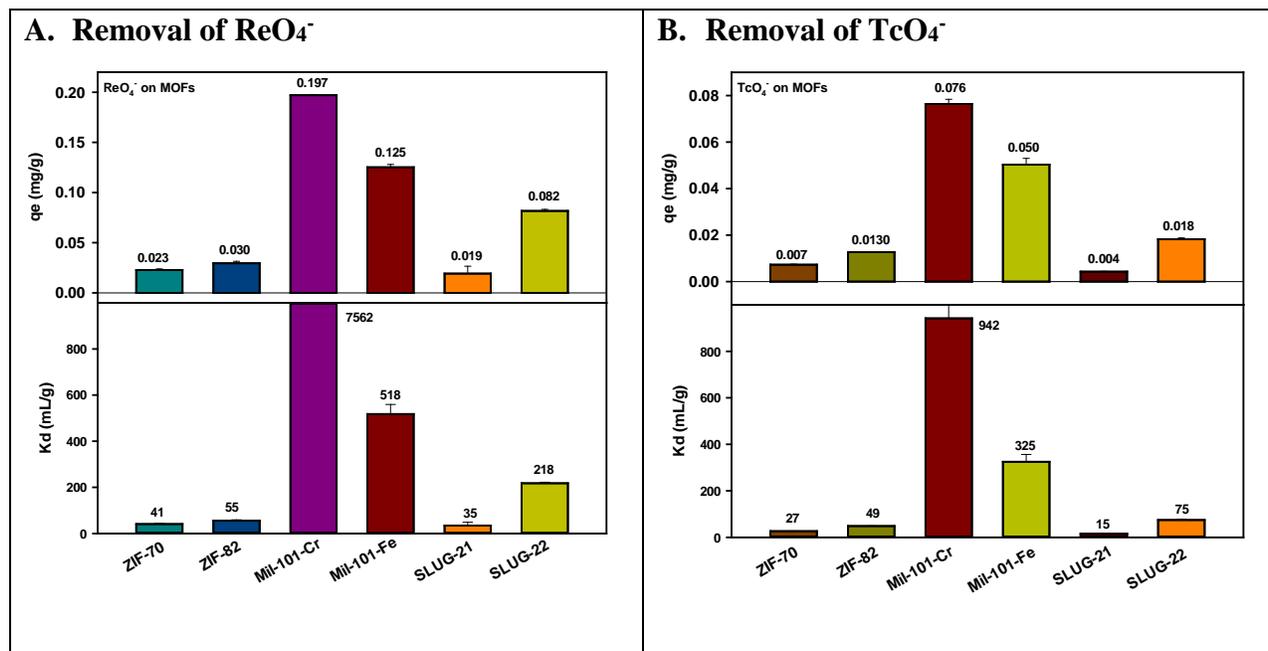
where  $C_0$  and  $C_e$  were Tc or Re concentrations before and after adsorption, respectively,  $V$  was the total volume of liquid phase (i.e., groundwater), and  $M$  was the mass of the solid material.

The adsorption coefficients of ZIF-8, ZIF-67, ZIF-70 and UiO-66- $\text{NO}_2$  from the first-round sample preparation are summarized in Table 1. The results demonstrated that UiO-66- $\text{NO}_2$  was effective for sequestering  $\text{TcO}_4^-$  from groundwater at pH 3 and 7.5. Some MOFs were not stable under acid conditions. New MOFs need to be developed for Tc uptake under alkaline conditions.

**Table 1 Adsorption coefficient ( $K_d$ , mL/g) of ZIF-8, ZIF-67, ZIF-70 and UiO-66-NO<sub>2</sub> for TcO<sub>4</sub><sup>-</sup> removal from artificial groundwater of different pH values**

MOFs	$K_d$ (mL/g)		
	pH 3	pH 7.5	pH 12
ZIF-8	unstable	72	0
ZIF-67	unstable	9	0
ZIF-70	12	0	0
UiO-66-NO <sub>2</sub>	750	156	0

The adsorption quantities of ZIF-70, ZIF-82, MIL-101(Cr), MIL-101(Fe), SLUG-21 and SLUG-22 for TcO<sub>4</sub><sup>-</sup> sequestration from artificial groundwater at equilibration pH values were determined using batch experiments. The results (Figure 2) demonstrated that MIL-101(Cr) was most effective for TcO<sub>4</sub><sup>-</sup> and ReO<sub>4</sub><sup>-</sup> sequestration from artificial groundwater at equilibration pH values with the removal capacity of 0.076 mg/g for Tc and 0.197 mg/g for Re. MIL-101(Fe) and SLUG-22 were moderately effective for TcO<sub>4</sub><sup>-</sup> and ReO<sub>4</sub><sup>-</sup> sequestration under similar conditions.

**Figure 2. Removal capacity of MOFs for ReO<sub>4</sub><sup>-</sup> (A) and TcO<sub>4</sub><sup>-</sup> (B) sequestration from artificial groundwater.**

#### 4. Potential for economical vitrification

Thermogravimetric analyses of selected MOFs were conducted (Figure 1). The results indicated that these MOFs may be vitrified using economic methods, like heating under 400 °C.

The actual vitrification technology will be developed for the eventually selected MOFs that are demonstrated to have high capacity for  $\text{TcO}_4^-$  removal and improved stability in aqueous media.

### FY2018 Accomplishments

- Ten MOFs of known crystal structure and chemistry were synthesized, characterized and evaluated for  $\text{TcO}_4^-$  and  $\text{ReO}_4^-$  sequestration from artificial groundwater.
- UiO-66- $\text{NO}_2$  was effective for sequestering  $\text{TcO}_4^-$  from groundwater at pH 3 and 7.5.
- MIL-101(Cr) was most effective for  $\text{TcO}_4^-$  and  $\text{ReO}_4^-$  sequestration from artificial groundwater at equilibration pHs with the removal capacity of 0.076 mg/g for Tc and 0.197 mg/g for Re. MIL-101(Fe) and SLUG-22 were moderately effective for  $\text{TcO}_4^-$  and  $\text{ReO}_4^-$  sequestration under similar conditions.
- MIL-101(Cr) is targeted as a top candidate for developing new MOFs with higher  $\text{TcO}_4^-$  removal capacity and improved stability during FY19. MIL-101-Fe, UiO-66- $\text{NO}_2$  and SLUG-22 are potentially candidates.

### Future Directions

- Develop new MOFs for higher  $\text{TcO}_4^-$  removal capacity and improved stability, especially under alkaline conditions.
- Evaluate MOFs for  $\text{TcO}_4^-$  removal under different environmental and geochemical conditions (i.e., pH, Eh, solution chemistry) and for  $\text{TcO}_4^-$  separation from liquid nuclear wastes.
- Prepare proposals to DOE EM Soil & Groundwater Remediation Program, International Program, and DOE Nuclear Energy program.

### FY 2018 Publications/Presentations

**Dien Li**, Natalia B. Shustova, Kathryn M. Taylor-Pashow, Daniel I. Kaplan, Jake W. Amoroso, Thomas D. Bennett, Metal-Organic-Frameworks (MOFs) for Sequestration and Stabilization of Pertechnetate ( $\text{TcO}_4^-$ ) from Groundwater and Liquid Nuclear Wastes, *International Conference of Heavy Metals in the Environment 2018*, Athens, USA, July 21-25, **2018**.

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## Acronyms

DOE	Department of Energy
MIL	Materials Institute Lavoisier
SLUG	University of California, Santa Cruz
SRS	Savannah River Site
TGA	Thermogravimetric analysis
UiO	University of Oslo
XRD	X-ray diffraction
ZIF	Zeolitic Imidazolate Framework

## Intellectual Property

N/A

## Total Number of Post-Doctoral Researchers

0.5 of post-doctoral researcher involved through the subcontractor with the University of South Carolina.