



ABSORPTION OF CARBON DIOXIDE IN WASTE TANKS (U)

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This is a Technical Report

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TECHNICAL DIVISION
SAVANNAH RIVER LABORATORY

Bicarbonate, Carbonate,
Hydroxide Depletion, Model,
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MEMORANDUM

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ABSORPTION OF CARBON DIOXIDE IN WASTE TANKS

SUMMARY

Air flow rates and carbon dioxide concentrations of air entering and exiting eight H-Area waste tanks were monitored for a period of one year. The average instantaneous concentration of carbon dioxide in air is within the range reported offsite, and therefore is not affected by operation of the coal-fired power plant adjacent to the tank farm. Waste solutions in each of the tanks were observed to be continuously absorbing carbon dioxide.

The rate of absorption of carbon dioxide decreased linearly with the pH of the solution. Personnel exposure associated with the routine sampling and analysis of radioactive wastes stored at *SRP Site* to determine the levels of corrosion inhibitors in solution could be reduced by monitoring the absorption of carbon dioxide and using the relationship between pH and carbon dioxide absorption to determine the free hydroxide concentration in solution.

Based on the average instantaneous concentration of carbon dioxide in air and the fraction of carbon dioxide absorbed in Tank 42H, the rate of change in the pH and the total carbonate concentration of

the washed sludge supernate has been calculated. Good agreement between experimental and calculated data was found, verifying the previously reported model for calculating the rate of hydroxide depletion in salt solutions.

INTRODUCTION

Full-scale demonstrations of the aluminum dissolution and washing of sludge and the decontamination of salt solution were conducted in 1983. Since the demonstrations, the washed sludge and washed precipitate have been stored in Tanks 42H and 48H, respectively. After storage for approximately twenty months, the free hydroxide level in both tanks had fallen below the minimum level of 0.01 M established by the current technical standard.¹ At that time, additional sodium hydroxide was added to the storage tank to raise the level to within the technical standard.²

Because of the increased potential for tank corrosion and the possibility of significantly higher sodium levels in the melter feed as the result of adding sodium hydroxide³, a program was initiated to determine the mechanism and rate of hydroxide depletion and to develop a model for calculating the time dependent concentrations of inhibitors in solution.⁴

EXPERIMENTAL

Carbon dioxide levels were determined using a Foxboro Miran™ 101 portable gas analyzer. The analyzer was zeroed with purified nitrogen and calibrated with two different standard carbon dioxide/air gas mixtures prior to each weekly measurement. The instantaneous atmospheric carbon dioxide level was measured at each tank. The analyzer was then placed in contact with the air being exhausted from the vent stack via plastic tubing. After 1-2 minutes a constant, instantaneous carbon dioxide reading was recorded. Flow rates were recorded using the flowmeter gauges located at each tank. The flow rate of Tank 42H was determined using a fan-type anemometer.

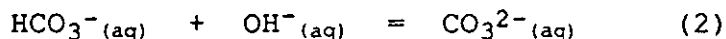
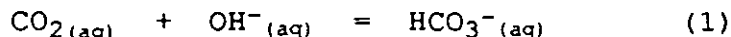
Dip samples from Tank 42H were taken by Waste Management Operations and transported to SRL. Upon arrival at SRL, the samples were placed into the high level cells. Each sample was well mixed prior to opening the sample bottle. After opening, the mixture was filtered through a Nalgene™ 0.45 µm disposable filter. The solids were discarded. The filtrate was then analyzed to determine inhibitor levels. During processing of the dip samples, exposure of the liquid to air was minimized to limit the absorption of carbon dioxide. Similar procedures were used by the F-Area Laboratory to determine inhibitor levels in waste solutions stored in the other H-Area Tanks.

Weather data were obtained from the SRL Weather Center. Statistical analyses of the experimental data were performed on a MacIntosh™ Plus personal computer using the StatView 512+™ software package.

DISCUSSION

Absorption Mechanism

Initial experiments indicated that as the concentration of hydroxide in solution decreased, the concentration of carbonate increased.⁵ This result suggested that the hydroxide in the waste solutions was being depleted by the absorption and reaction of atmospheric carbon dioxide. Alkaline solutions are known to readily absorb carbon dioxide forming bicarbonate and carbonate species as in equations 1 and 2.



Since air contains carbon dioxide and is continuously swept over the waste solutions to prevent the buildup of explosive mixtures of hydrogen, there is a continuous source of carbon dioxide for absorption. According to this mechanism, each tank would absorb carbon dioxide until steady-state conditions are reached as established by hydroxide, bicarbonate, and carbonate equilibria in solution and carbon dioxide concentrations in air.

To verify this mechanism and provide experimental data to model the rate of hydroxide depletion, the flow rates of air through the waste tanks, the levels of hydroxide and carbonate in the waste solutions, and the carbon dioxide concentrations in the air entering and exiting, waste tanks were measured on a regular basis. During this time, a test was conducted to allow the pH of the washed sludge supernate in Tank 42H to fall below technical standards.⁶

Measured Carbon Dioxide Levels

Atmospheric carbon dioxide concentrations entering and exiting eight waste tanks in H-Area were monitored on a weekly basis from January 23, 1986 to January 20, 1987. Tanks monitored included 38H, 39H, 41H, 42H, 43H, 48H, 49H, and 50H. Tanks 38H and 43H contain LHW and are feed tanks for the evaporator. Tank 41H contains evaporator concentrates and Tank 39H contains HHW. Washed precipitate, spent wash water, and decontaminated supernate are stored in tanks 48H, 49H, and 50H, respectively. Until September 15, 1986, only washed sludge was stored in Tank 42H. At that time, unwashed sludge from Tank 18F was transferred into Tank 42H and mixed with the washed sludge.

The average instantaneous atmospheric concentration of carbon dioxide during the monitoring period was 361 ± 19 ppm. The carbon dioxide concentrations ranged from a low of 325 ppm to a high of 440 ppm. These results are consistent with values reported in the literature.⁷

Concentrations were measured during daylight and after sunset in a variety of weather and seasonal conditions. No statistically significant variations in carbon dioxide concentrations were observed with respect to time, temperature, wind direction, wind speed, and season (see Tables I and II). Higher average concentrations (ca. 15 ppm) were observed for the fall and winter seasons as compared to spring and summer. This variation is within the normal seasonal variations for atmospheric carbon dioxide concentrations.

It was believed that operation of the coal-fired power plant adjacent to the waste tanks could significantly alter the atmospheric carbon dioxide concentrations.⁸ However, this was not observed. The average concentration downwind of the power plant was 361 ± 16 ppm as compared to 360 ± 20 ppm upwind.

Carbon Dioxide Absorption

The average air flow rate, atmospheric carbon dioxide concentration, and absorption fraction for each tank is presented in Table III. The absorption fraction is calculated by dividing the difference of the carbon dioxide concentrations entering and exiting the tank by the carbon dioxide concentration entering the tank (equation 3).

$$\text{Absorption Fraction} = \{[\text{CO}_2]_{\text{enter}} - [\text{CO}_2]_{\text{exit}}\} / [\text{CO}_2]_{\text{enter}} \quad (3)$$

Tanks 38H, 39H, 41H, 43H, 49H, and 50H showed no significant changes in the measured absorption of carbon dioxide during the monitoring period. Approximately 77% of the carbon dioxide entering these tanks was absorbed.

In contrast, the absorption fraction in Tanks 42H and 48H was observed to change significantly. Tank 42H showed a very low absorption fraction during the first four months followed by an abrupt increase. The absorption fraction remained at the higher value during the remaining eight months. Tank 48H showed a high absorption fraction during the first eight months of monitoring. The absorption fraction decreased steadily during the final four months.

The changes in the rate of carbon dioxide absorption in these two tanks are due to changes in the concentration of free hydroxide in the liquid phase. From January through April, the free hydroxide level in the Tank 42H supernate was approximately 1.0×10^{-4} molar (pH 10). At this low free hydroxide level, the thermodynamics are less favorable for carbon dioxide absorption, and, thus, the absorption is low. In early May, additional sodium hydroxide was added to Tank 42H to raise the free hydroxide level to 0.0327 molar (pH 12.5). At the increased free hydroxide level, the carbon dioxide absorption is more favorable and the observed absorption fraction was increased by a factor of about six (see Table III).

In the case of Tank 48H, the decrease in the absorption fraction was accompanied by a decrease in the free hydroxide concentration from 0.12 M to 0.064 M. The lower free hydroxide level resulted in lower absorption rates in the waste tank.

Linear regression analysis of the data was carried out to determine if there existed any correlations between any of the measured variables and the absorption fraction. Table IV shows a list of selected coefficients of correlation. Based on the low coefficients, the rate of absorption or absorption fraction did not correlate with time-of-day, date, atmospheric carbon dioxide concentration, wind direction, wind speed, or temperature. Higher coefficients were observed for nitrate concentration, specific gravity, and pH.

From reaction kinetics and thermodynamics, one would expect the air flow rate and the carbon dioxide concentration to significantly affect the absorption. However, in the case of these data, the range of both of these variables is limited and statistical correlation may not be observed.

The highest coefficient observed was between the pH of the solution and the absorption fraction for each tank. The pH was calculated from analytically determined hydroxide ion concentrations using equation 4.

$$\text{pH} = 14 + \log [\text{OH}^-] \quad (4)$$

Figure 1 shows a plot of the average pH of the solution in each tank versus the average absorption fraction measured for each of the waste tanks. The solid line represents the linear least squares fit of the data and had the form;

$$y = 0.165x - 1.569, \quad (5)$$

where $x = \text{pH}$ and $y = \text{absorption fraction}$. The correlation coefficient for the linear regression analysis was 0.948. The dash lines represent the 95% confidence bands about the mean of the absorption fraction.

Determination of Hydroxide Levels by Carbon Dioxide Monitoring

The linear correlation between the pH of the solution and the absorption fraction may have practical application in monitoring inhibitor levels in the waste tanks. Currently, the hydroxide concentration in waste solutions is determined by sampling the contents of the waste tank and determining the concentration by titration. Alternatively, one could measure the fraction of carbon dioxide absorbed and use equation 6 to calculate the hydroxide concentration. Equation 6 is obtained by substituting equation 4 into equation 5 for the pH and rearranging to give an expression for the molar hydroxide concentration in terms of the absorption fraction.

$$[\text{OH}^-] = 10^{[(\text{Absorption Fraction} - 0.741)/0.165]} \quad (6)$$

supernate were calculated as a function of time using the carbon dioxide absorption model.

Figures 3 and 4 show the calculated pH and Figures 5 and 6 show the calculated carbonate concentrations, respectively, for the time periods, December, 1985 to May, 1985 and May, 1985 to March, 1986. Experimentally determined pH and carbonate levels of supernate samples taken from Tank 42H during these times are also plotted.

Two calculated curves are plotted for each case. These represent the minimum and maximum rates at which the pH and carbonate levels are reached and are based on the standard deviation of the experimentally measured carbon dioxide level, air flow rate, and absorption fraction.

There was good agreement between the observed and calculated pH and carbonate values for both of these time periods verifying that the carbon dioxide absorption model can be used to calculate the time dependent concentrations of hydroxide and carbonate in waste solutions stored in SRP waste tanks. The differences between the experimental and calculated values can be attributed to experimental errors in sampling and analytical techniques and assuming constant flow rates and absorption fractions during these times periods.

The pH decreased in a normal titration-like curve for a strong base for both the observed and calculated cases. The lower limit at pH 9.5 - 10.0 is consistent with formation of a bicarbonate/carbonate buffer system. The steady-state concentration of total carbonate is dependent on the initial hydroxide concentration and the partial pressure of the carbon dioxide in the air.

The total carbonate concentration increased linearly as expected by the mass-transfer limited absorption of carbon dioxide. During the range of pH values spanned, the predominant carbonate species changes from carbonate (high pH) to bicarbonate (low pH).

QUALITY ASSURANCE

The experimental results referenced in this memorandum were obtained in accordance with the Savannah River Laboratory Quality Assurance Program and Defense Waste Section Quality Assurance Plan.⁹ Experimental procedures and data are recorded in laboratory notebooks DPSTN #4379 and #4461 maintained by D. T. Hobbs.

If the absorption fraction indicated that the hydroxide concentration was close to the lower limit of the technical standard, a supernate sample would then be taken to confirm the hydroxide concentration. Personnel exposure would be significantly reduced by decreasing the number of liquid samples.

In addition to lower personnel exposure, the carbon dioxide monitoring instrument is portable and the technique is rapid. The ambient air analyzer used in this study weighed approximately eighteen pounds including the rechargeable battery pack and was easily operated by a single individual. Calibration of the instrument and sampling of the eight waste tanks routinely took less than one hour to complete.

Absorption Model

Because of the increased potential for corrosion of the carbon steel waste tank as free hydroxide is depleted, a model of the waste tanks was developed that would allow one to calculate the time dependent concentrations of hydroxide and carbonate species. A schematic representation of the model is shown in Figure 2.

The contents of each waste tank are assumed to be comprised of three phases, a gas phase, a liquid phase, and a solid phase. The gas phase is predominantly air with other volatile compounds that may be present (eg. benzene in washed precipitate slurry). The liquid phase consists of the sodium salts of inorganic and organic anions. The solid phase is comprised of inert water insoluble species which do not affect thermodynamic equilibria in either of the other two phases.

Air is swept over and contacts the liquid in the waste tank. Carbon dioxide in the air is absorbed by the liquid phase and reacts instantaneously with hydroxide to form bicarbonate (equation 1). If sufficient hydroxide is present, the bicarbonate reacts with additional hydroxide to form carbonate (equation 2). Carbon dioxide is continually absorbed until the partial pressure of carbon dioxide in equilibrium with the liquid phase equals the partial pressure of carbon dioxide in the air entering the tank.

The kinetics of reactions between hydroxide and carbon dioxide or bicarbonate are extremely fast. Thus, the rate determining step in depleting hydroxide in the liquid phase is the transfer of carbon dioxide across the gas-liquid interface.⁴ The rate of carbon dioxide absorbed is calculated by multiplying the air flow rate through the tank by the concentration of carbon dioxide in air and by the fraction of carbon dioxide absorbed (equation 7).

$$\text{Rate} = \text{Flow Rate} \times [\text{CO}_2]_{\text{air}} \times \text{Absorption Fraction} \quad (7)$$

Assuming a constant air flowrate through Tank 42H at 430 ($\pm 10\%$) scfm, using the average atmospheric carbon dioxide concentration (361 ± 19 ppm), and assuming that an absorption fraction of 0.581 ± 0.039 , the pH and the total carbonate concentrations in Tank 42H

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2. DPSTS-241-5.01, "Limitations on Waste Tank Contents".
3. D. T. Hobbs, DPST-85-861, "Hydroxide Depletion in Dilute Supernates Stored in Waste Tanks", October 10, 1985.
4. D. T. Hobbs and R. M. Wallace, DPST-85-846, "Hydroxide Depletion in Waste Storage Tanks by Reaction with Carbon Dioxide", October 14, 1986.
5. D. T. Hobbs, September Monthly Status Report-Interim Waste Technology Division, October 7, 1985.
6. P. D. D'Entremont, Technical Authorization No. 2-1099, "Monitoring Hydroxide Depletion of Dilute Supernate in Tank 42".
7. J. M. Wallace and P. U. Hobbs, "Atmospheric Science: An Introductory Survey", Academic Press, New York, 1977, p. 10.
8. D. T. Hobbs, DPST-86-735, "Absorption of Carbon Dioxide in Waste Tanks", October 21, 1986.
9. Savannah River Laboratory Quality Assurance Manual, DPSTQA-83-4-2, November 1983.

Table I

ATMOSPHERIC CARBON DIOXIDE CONCENTRATIONS

<u>Season</u>	<u>Average Concentration</u> (ppm)
Spring (March-May)	347 \pm 7
Summer (June-August)	352 \pm 17
Fall (September-November)	363 \pm 18
Winter (December-February)	370 \pm 19
All Seasons	361 \pm 19

TABLE II

**LINEAR REGRESSION ANALYSIS OF ATMOSPHERIC
CARBON DIOXIDE CONCENTRATION**

X	Y	<u>CORRELATION COEFFICIENT</u>
Date	CO ₂ Concentration	0.339
Time	CO ₂ Concentration	0.123
Wind Direction	CO ₂ Concentration	0.230
Wind Speed	CO ₂ Concentration	0.099
Temperature	CO ₂ Concentration	0.496

TABLE III

**AVERAGE AIR FLOW RATES AND CARBON DIOXIDE ABSORPTION
FRACTIONS FOR H-AREA WASTE TANKS**

TANK	COMPOSITION	FLOWRATE (scfm)	[CO ₂] ENTER (ppm)	ABSORPTION FRACTION
50	Decontaminated Supernate	293 ± 35	362 ± 18	0.746 ± 0.037
49	Spent Wash Water	336 ± 11	357 ± 14	0.697 ± 0.050
48	Washed Precipitate			
	- High [OH]	266 ± 44	350 ± 14	0.715 ± 0.046
	- Low [OH]	231 ± 26	361 ± 15	0.400 ± 0.055
43	LHW	311 ± 60	361 ± 19	0.834 ± 0.056
42	Washed Sludge			
	- Low [OH]	430	355 ± 22	0.096 ± 0.023
	- High [OH]	430	364 ± 21	0.581 ± 0.039
41	Evaporator Conc.	351 ± 21	366 ± 22	0.756 ± 0.065
39	HHW	264 ± 62	360 ± 20	0.791 ± 0.069
38	LHW	268 ± 40	359 ± 18	0.798 ± 0.052
All	-	290 ± 54	361 ± 19	0.695 ± 0.152
All except 42 and 48		299 ± 53	361 ± 19	0.770 ± 0.070

TABLE IV

LINEAR REGRESSION ANALYSIS OF ABSORPTION FRACTION

<u>X</u>	<u>Y</u>	<u>CORRELATION COEFFICIENT</u>
Date	Absorption Fraction	0.037
Time	Absorption Fraction	-0.094
Flow Rate	Absorption Fraction	0.273
CO ₂ Conc.	Absorption Fraction	-0.087
Wind Direction	Absorption Fraction	-0.036
Wind Speed	Absorption Fraction	4.8×10^{-4}
Temperature	Absorption Fraction	0.023
Nitrate Concentration	Absorption Fraction	0.641
Specific Gravity	Absorption Fraction	0.554
pH	Absorption Fraction	0.948

FIGURE 1. Plot of Average Absorption Fraction Versus Average Calculated pH

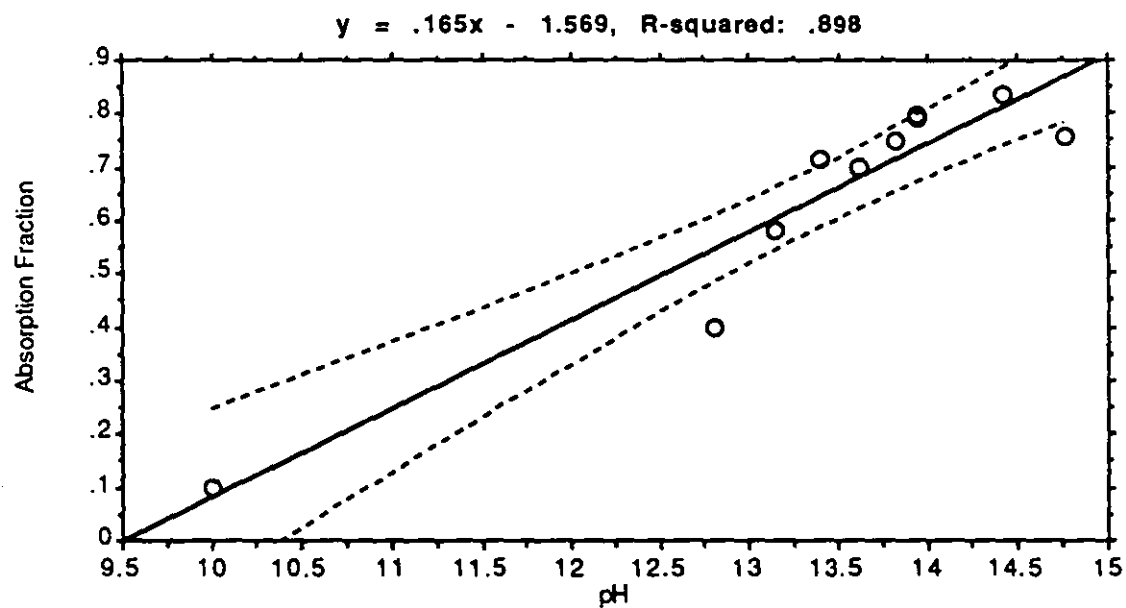
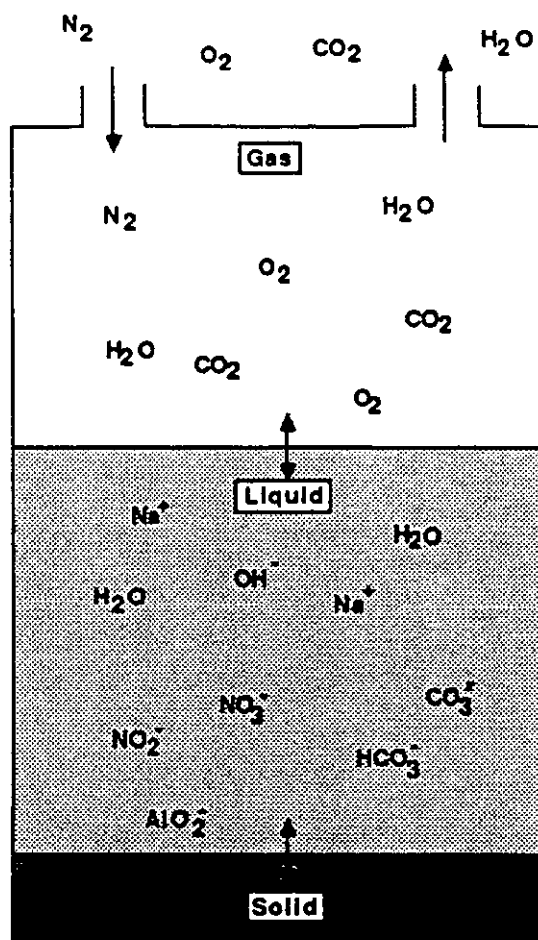


FIGURE 2. Schematic Representation of Waste Tank Absorption Model



$$\text{Steady-State : } p(CO_2)^{\text{air}} = p(CO_2)^{\text{tank}}$$

FIGURE 3. Calculated and Observed pH of Washed Sludge Supernate from December, 1984 to May, 1985.

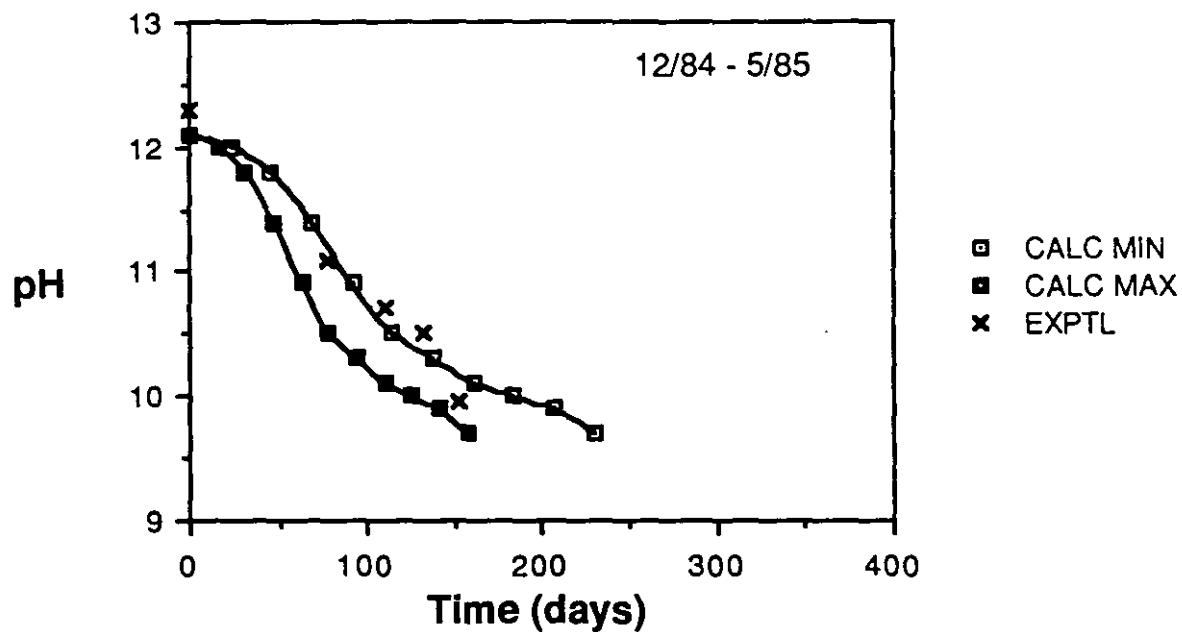


FIGURE 4. Calculated and Observed pH of Washed Sludge Supernate from May, 1985 to March, 1986.

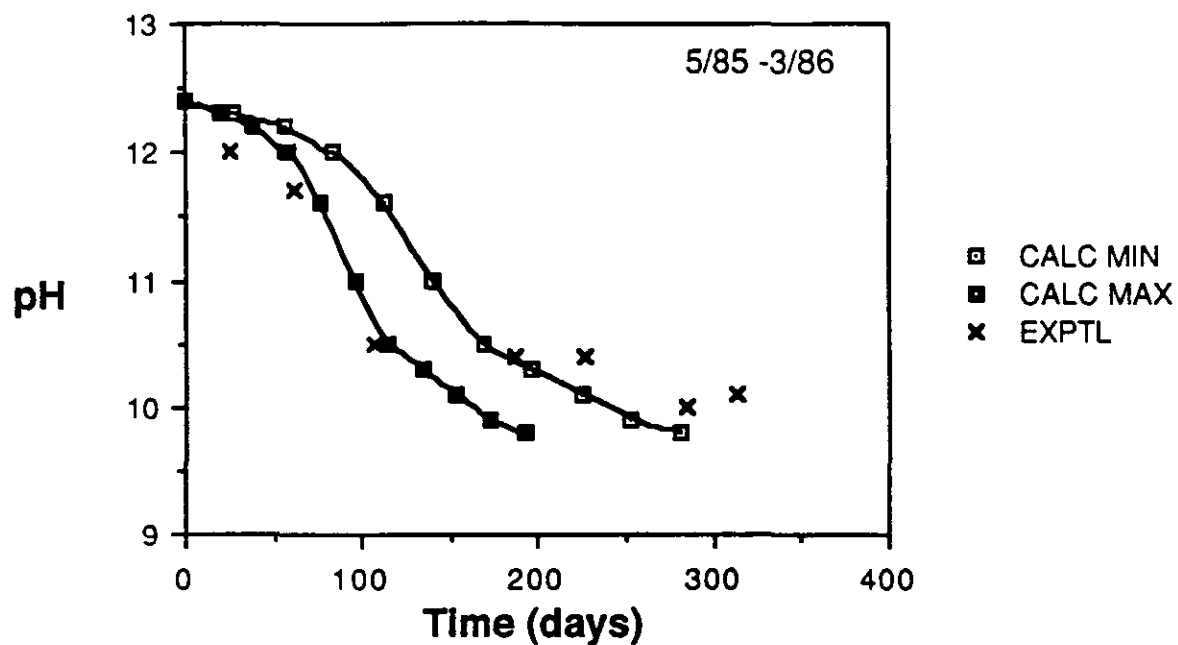


FIGURE 5. Calculated and Observed Total Carbonate in Washed Sludge Supernate from December, 1984 to May, 1985.

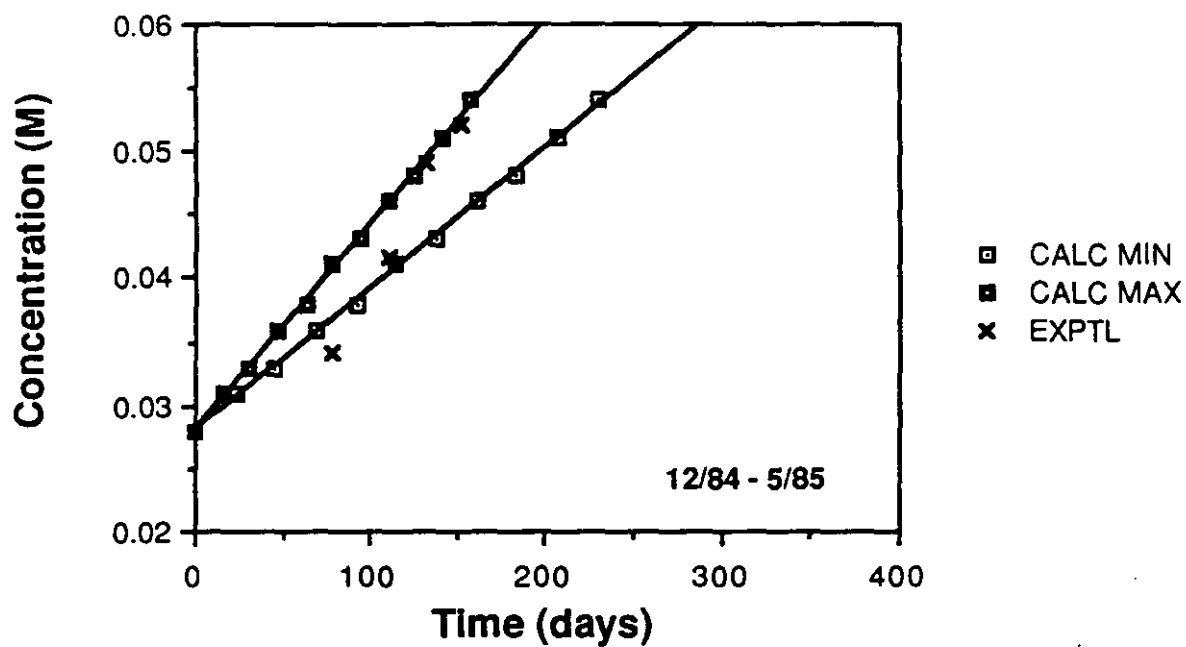


FIGURE 6. Calculated and Observed Total Carbonate in Washed Sludge Supernate from May, 1985 to March, 1986.

