



LYSIMETER TESTS OF SRP WASTE FORMS

R. L. HOOKER
R. W. ROOT, JR.

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ABSTRACT

A field study, estimated to last 10 years, has been started to define leaching and migration rates of radionuclides from typical SRP buried wastes. The study utilizes 42 lysimeters (6-ft or 10-ft diameter by 10-ft deep) which have been charged with soil and waste to simulate burial ground conditions. Eight waste forms were selected for the study, which represent the bulk of the wastes generated at SRP. This report describes the lysimeter design, the physical and radiological characteristics of the wastes, and the experimental approach. Calculations have also been made which predict the migration of various radionuclides in the lysimeter soil. The calculations should provide guidance during the course of the study, and are the basis of recommendations made for collecting and interpreting data so that important parameters of migration can be evaluated.

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INTRODUCTION

The Savannah River Plant (SRP) site occupies an approximately circular area (192,000 acres) in South Carolina bounded on the southwest by the Savannah River and centered approximately 25 miles southeast of Augusta, Georgia. Solid radioactive defense wastes are stored in one centrally located burial ground. The original 76-acre burial ground was placed in operation in 1953 and 119 acres were added in 1972.^{1,2}

Studies at the Savannah River Laboratory (SRL) are currently investigating the long-term consequences of past and current burial ground operations. These studies involve radionuclide chemistry, trench sampling, groundwater modeling and monitoring, and waste migration.

A field study was recently begun which uses lysimeters to define leaching and migration rates of radionuclides from typical SRP buried wastes. A lysimeter measures water percolation through soils and determines the soluble constituents removed by drainage. Lysimeters are typically used in agricultural studies, but can be adapted for radionuclide migration studies.³ Lysimeters (6-ft or 10-ft diameter by 10-ft deep) were installed at SRP to function as miniature burial grounds in which the amount and location of the wastes are controlled and provisions are made for collecting and sampling percolating rain-water. The rate at which radioactivity is leached from waste and the rate the activity moves through soil are primary factors in evaluating the future impact of buried waste on the environment and on man.

The wastes buried in the lysimeters are typical of those generated by the wide variety of plant and laboratory operations. The wastes include irradiated reactor fuel housings, ion exchange resins, waste plastics and tools from the plutonium finishing line, and laboratory glassware. The wastes contain ^{238}Pu , ^{239}Pu , ^{60}Co , ^{137}Cs , ^{90}Sr , ^{14}C , and other miscellaneous fission and activation products. Leaching and migration of plutonium is of interest, even though no significant amount of plutonium is currently being buried, because about 4000 Ci of ^{238}Pu and 400 Ci of ^{239}Pu were buried unencapsulated in trenches during the early years of SRP burial ground operation.

The lysimeter program is expected to continue for approximately 10 years, during which time leachate water samples will be taken. For those lysimeters which show no radioactivity in the leachate after 10 years, soil samples can be taken by excavation or coring to determine the extent of leaching and migration. Results from the study will be used in a mathematical model to calculate the long-term potential dose-to-man from the buried wastes.

Lysimeter Design and Installation

The lysimeters used by SRL were constructed of corrugated aluminum pipe sections (12 gage) which were coated with asphalt and are normally used for culverts (Figure 1). The bottom of the lysimeter is sloped to drain the percolate water to a low point where it can be pumped out. Because a few of the first lysimeters installed leaked after being loaded with soil and waste, the remainder were provided with a 20-mil thick polyvinylchloride liner and tested to ensure a leak rate below 0.15 L/hr.

Because the lysimeter site is between the old and the new burial grounds, soil, weather, and atmospheric fallout should be similar to that encountered in the burial grounds (Figure 2). There are five lysimeters for each of eight waste forms, plus two lysimeters in which no waste has been placed. The latter two lysimeters are used as controls to determine if airborne contamination from the nearby burial ground, separations areas, or regional atmospheric fallout is influencing lysimeter sample results.

Figure 3 shows a layout view of a typical set of five lysimeters. One of the lysimeters is 10 ft in diameter and the other four are 6 ft in diameter. Holding tanks (6-ft diameter x 6-ft high) are provided to receive all of the water removed from the lysimeters which contain transuranic waste.

The soil placed in the lysimeters was selected as being typical SRP burial ground soil. It was accumulated and then mixed with large earth-moving equipment so that variations among lysimeters would be minimized. Samples of soil were taken and will be characterized. Soil core samples can also be taken without disturbing the waste by entry through the side of the lysimeters. The latter may be necessary if the radionuclide is retained strongly by the soil such that it is not detectable in water collected in the sump. The data obtained will be used to calculate leaching and migration rates from the emplaced wastes.

Waste Forms and Characteristics

Eight waste forms were selected that represent most of the wastes in the SRP burial ground (Table 1). The wastes come from multiple generation points on the plant and may contain a number of radionuclides in various forms. A comprehensive description of the waste forms, the quantity and method of measurement of the radioactivity present in each lysimeter, and the date of waste emplacement are given in the Appendix (Table A-1).

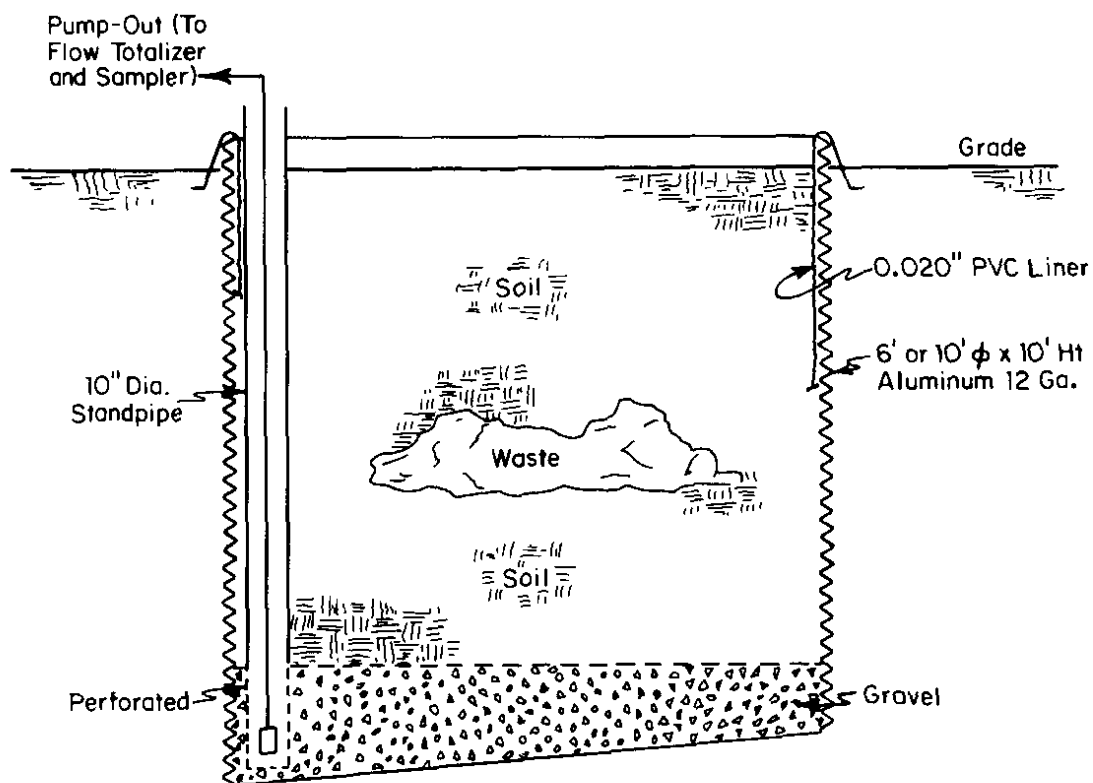


FIGURE 1. Lysimeter Cross Section

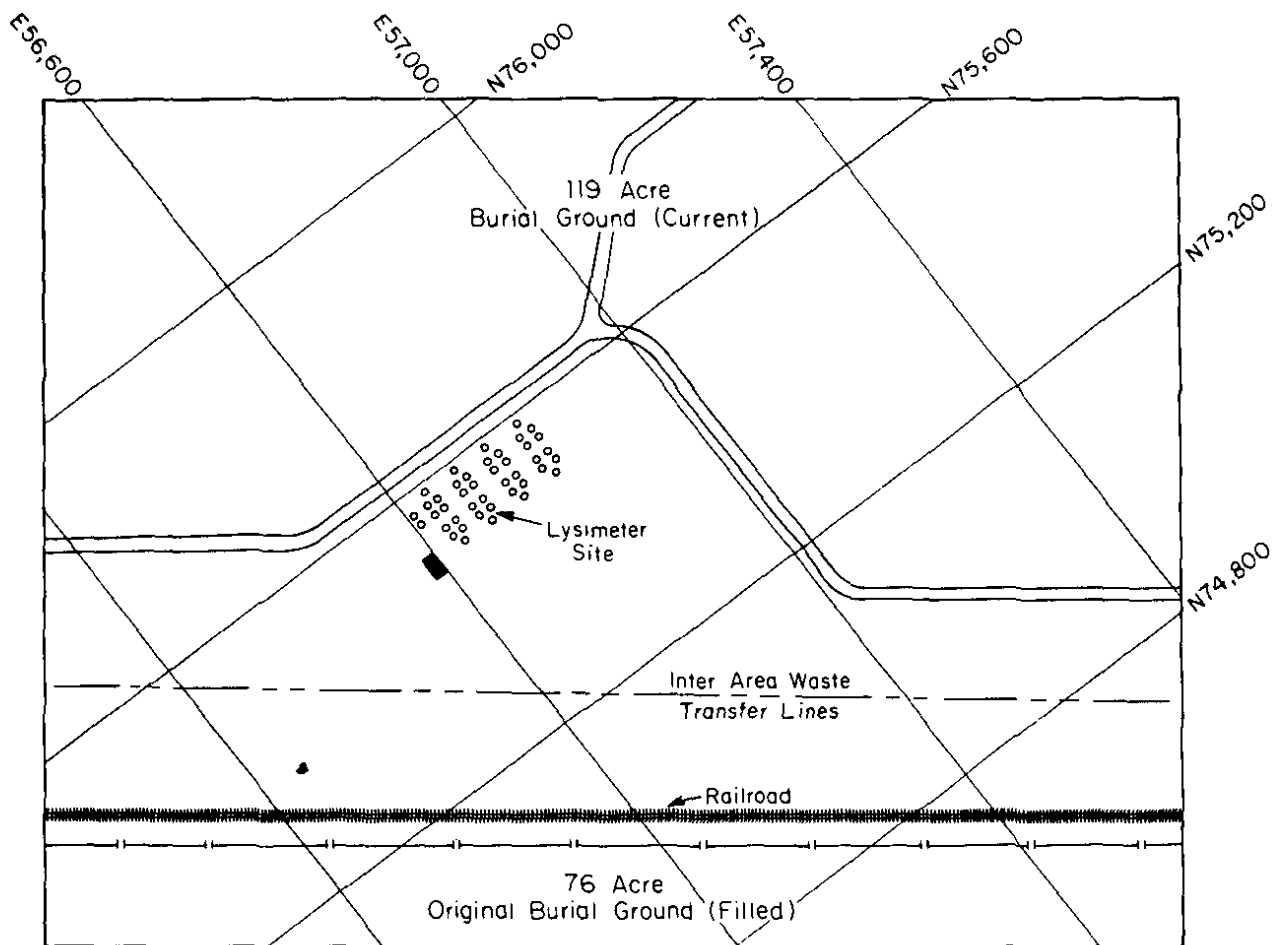
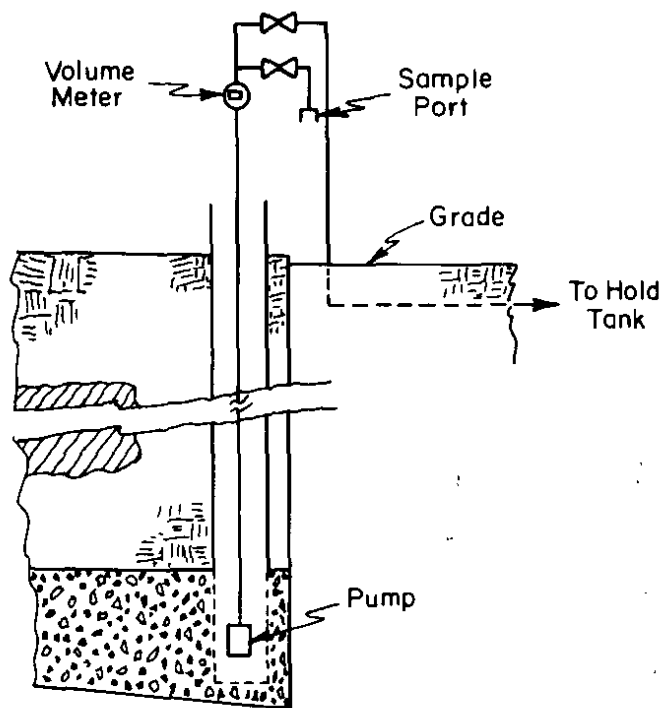
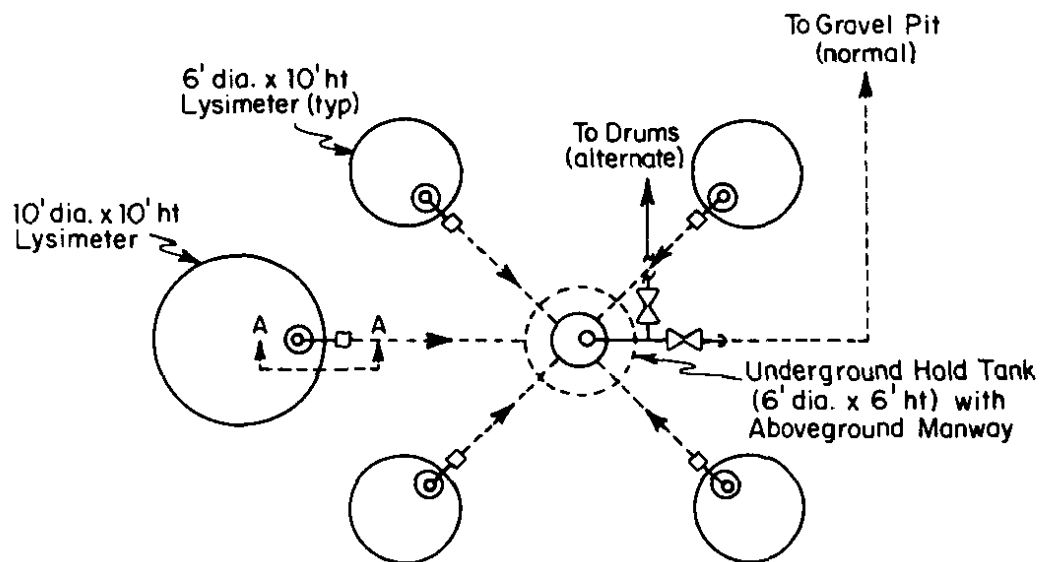


FIGURE 2. Plot Plan of Lysimeter Site



SECTION A-A (Typ)

FIGURE 3. Lysimeter Layout - Typical Set of Five

TABLE 1

Waste Types Buried in Lysimeters

| <u>Waste Forms</u> | <u>Primary Radioactivity Type</u> |
|--|--|
| Separations Laboratory Glove Box Waste | ^{238}Pu |
| Reactor Moderator Deionizer Resin | Fission products and activation products |
| Plutonium Finishing Line Job Control Waste | ^{239}Pu |
| SRL High Level Caves and Actinide Materials Facility (AMF) Waste | ^{238}Pu , fission products, and activation products |
| Separations Canyon Pipe Jumpers | ^{238}Pu , ^{239}Pu and a small amount of fission products |
| Separations Area Laboratory Waste | Fission products |
| Separations Area Canyon Job Control Waste | Fission products |
| Reactor Area Scrap Metal | Activation products |

Figures 4 through 10 are photographs of wastes taken during or after placement in lysimeters. Most of the views are through an opening in a plastic cover placed over the lysimeter to control airborne contamination. No photograph was taken of Separations Canyon Job Control waste, but its appearance is similar to Separations Laboratory Glove Box Waste.

Figure 11 shows the portable filter-blower which was fabricated for this program and used to control airborne contamination during charging of the lysimeter by inducing inflow of air through an opening in the plastic cover over the top of the lysimeter. The unit has two high efficiency particulate air (HEPA) filters in series. The design basis of the filter-blower was to maintain an inward air flow velocity of 150 ft/min through the opening in the plastic. The blower is powered with a 10 HP gasoline engine. Although the lysimeter loading operations were conducted outdoors in ambient air currents, spread of airborne contamination was successfully avoided with this apparatus.



FIGURE 4. Separations Laboratory Glove Box Waste in Lysimeter

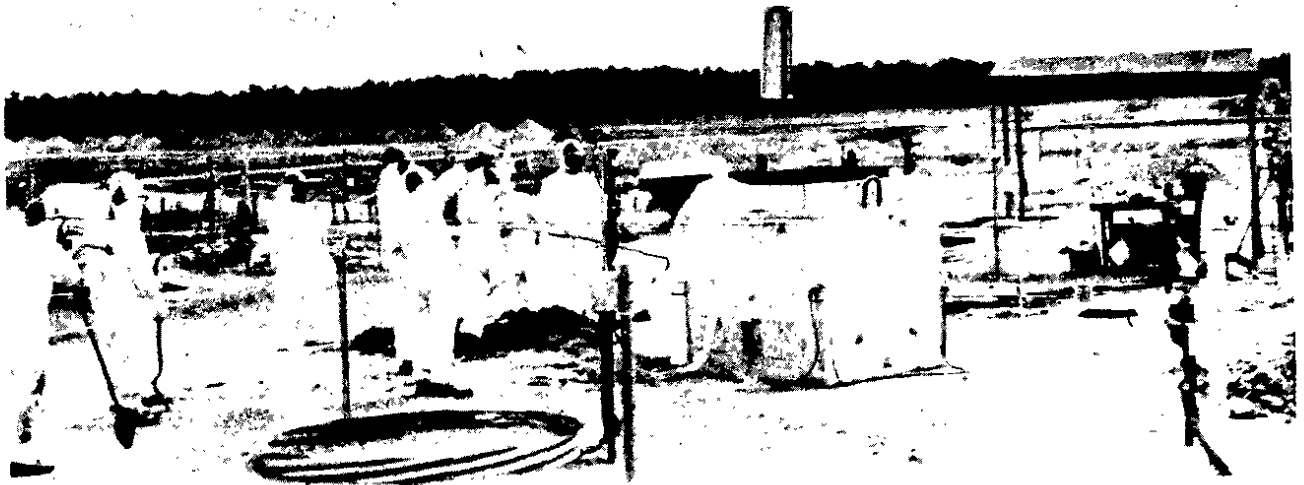


FIGURE 5. Reactor Moderator Resin Being Dumped into Lysimeter

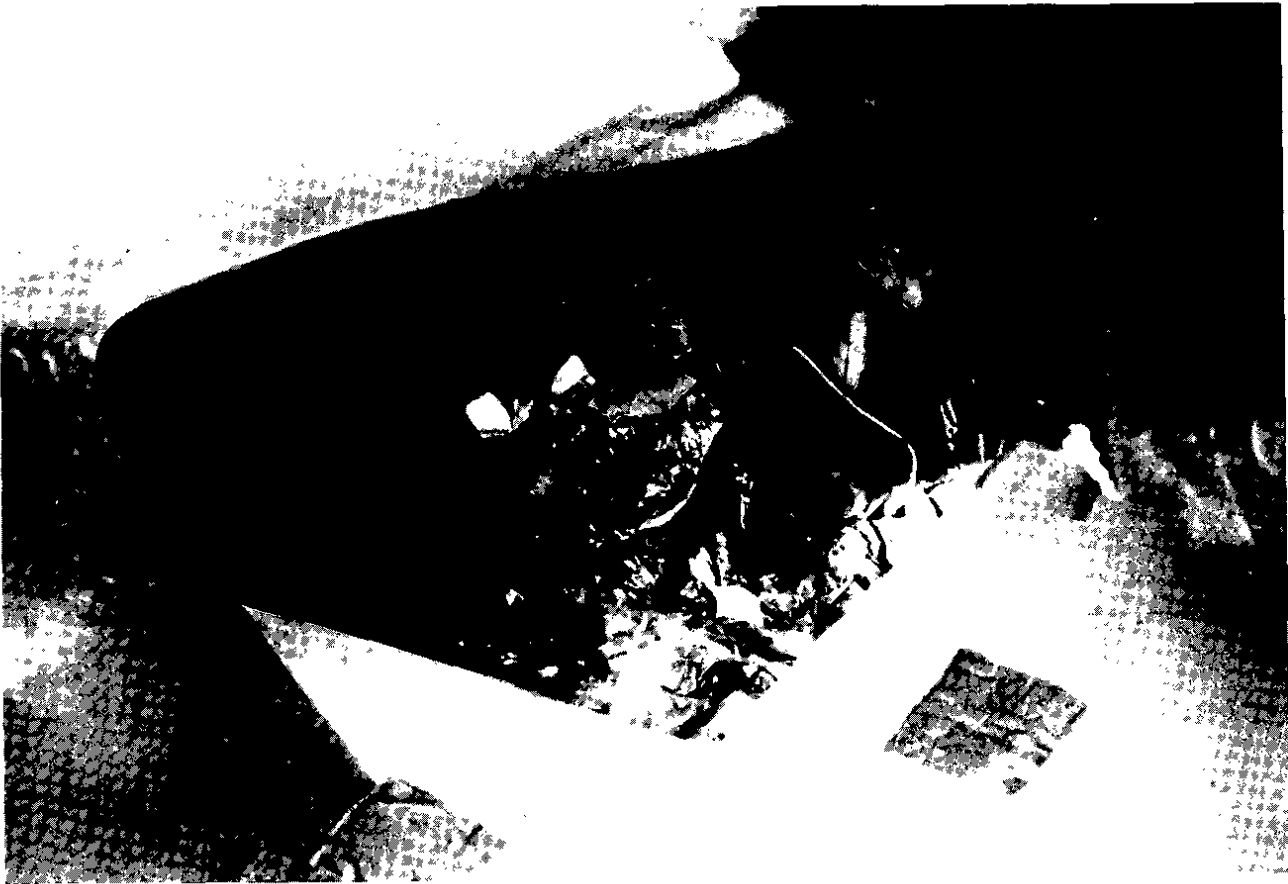


FIGURE 6. Plutonium Finishing Line Job Control Waste in Lysimeter



FIGURE 7. SRL High Level Caves Waste in Lysimeter



FIGURE 8. Separations Canyon Pipe Jumpers in Lysimeter

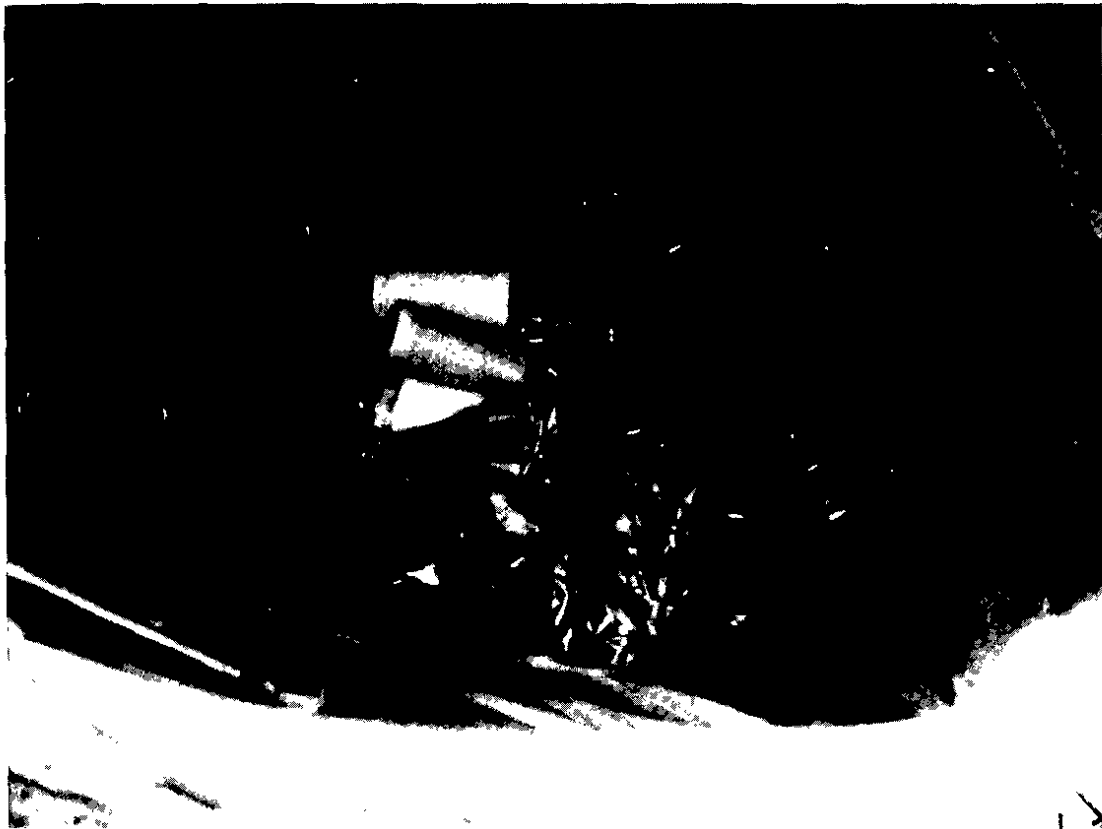


FIGURE 9. Separations Area Laboratory (Bldg. 772-F) Waste in Lysimeter



FIGURE 10. Reactor Area Scrap Metal in Lysimeter



FIGURE 11. Portable Filter-Blower Assembly

Figure 12 shows long-handled knives which were used for opening the waste packages which were placed in the lysimeters. Packages were intentionally breached to allow leaching to start uniformly soon after the lysimeters were loaded.

Lysimeter Operation

Two of the 6-ft diameter lysimeters of each set are operated in an "unsaturated" mode; i.e., with the water level maintained well below the waste zone. The other two 6-ft diameter lysimeters of each set are operated in a "saturated" mode; i.e., with the water level maintained above the waste zone. The saturated mode simulates a condition which sometimes exists in trenches, due to perched water at the interface between undisturbed and backfilled soil at the trench bottom. The 10-ft diameter lysimeter is operated in the unsaturated mode. This lysimeter was made large enough to support the root system of a pine tree. Pensacola bahia grass, the same vegetative cover used in the SRP burial ground to prevent erosion, is planted in all the lysimeters. The grass will be limed and fertilized at the same rate as the burial ground. Radionuclide uptake by the grass and by the pine trees will be measured under these conditions where the location, amount, and type of radionuclide are known.

Each lysimeter has a sump placed at the low point to remove percolate water (Figure 3). A pump is placed in the sump with its suction opening at the minimum desired water level. A meter indicates the volume of water removed. Periodic attendance of the lysimeter is required to maintain and confirm that the water levels do not exceed the intended levels, particularly following periods of heavy rainfall. Currently, the pumps are operated manually, but in the future the pumps will automatically turn on or off to maintain a water level within a 6-in. operating range. Water level indication is also planned to confirm that the pumps are operating satisfactorily. Water removed from the lysimeters is sampled and normally disposed of in gravel-lined pits. If radioactivity is detected above certain limits, the water will be collected and routed to a Separations Area for disposal.

Lysimeter Modeling

Modeling of radionuclide movement through the SRP lysimeters was done to help predict when measurable levels of activity might be seen in effluent water and thus provide guidance for long-term lysimeter operation. The parameters identified as important for transport through the lysimeters can be used to link laboratory measurements with actual radionuclide migration in the burial trenches.

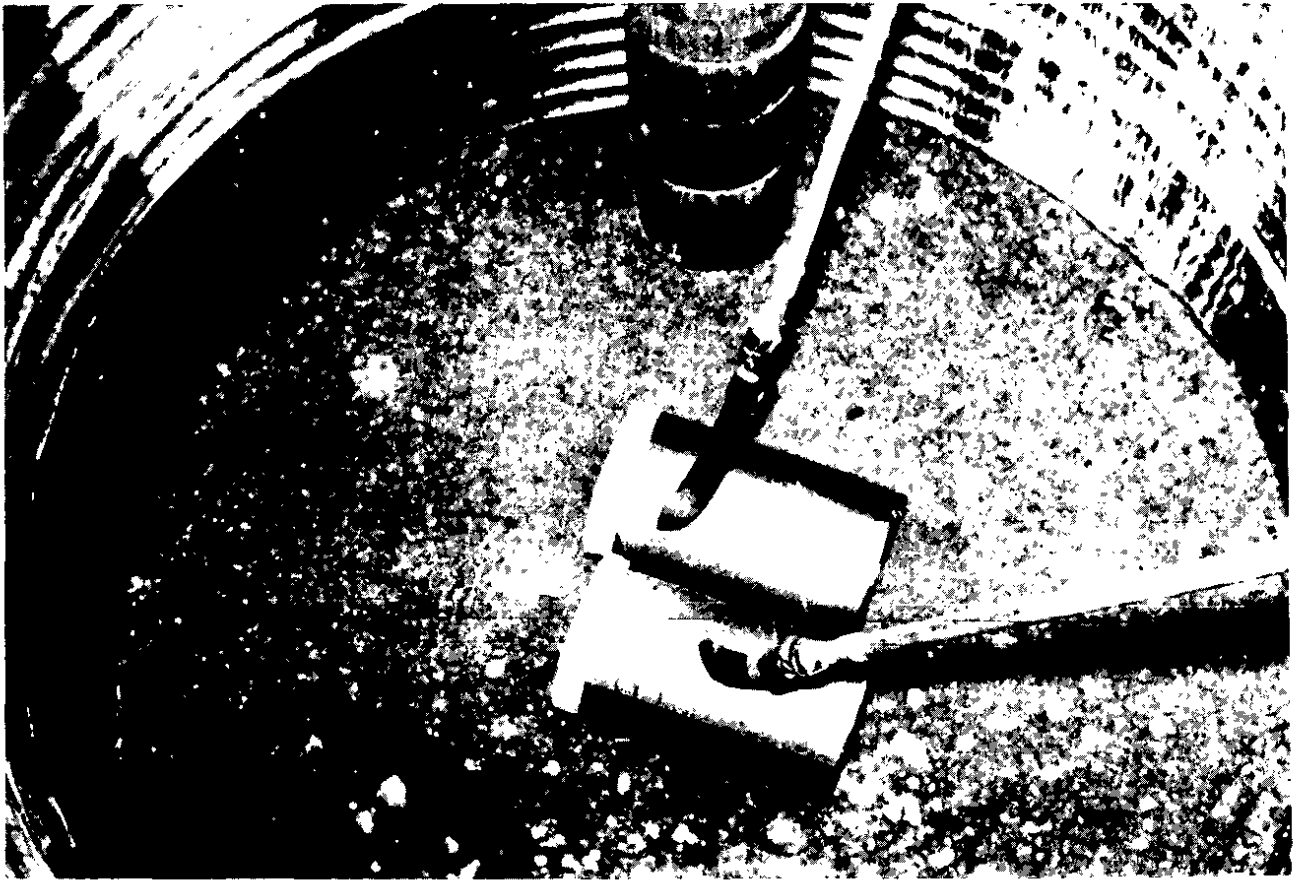


FIGURE 12. Long-Handled Knives Used to Open waste packages
in Lysimeter

Transport Equation

The 3-ft layer of soil beneath the waste zone (Figure 1) retards movement of most radionuclides by adsorption and ion exchange. The time of appearance and the concentration of a given radionuclide in the water collected at the bottom of the lysimeter will depend primarily on four factors: the leach rate, the water velocity, the distribution coefficient of the radionuclide between the soil and soil water, and the dispersion coefficient. The leach rate and rainfall combine to set the concentration (C_0) of the radionuclide in the soil water just below the waste.

Radionuclide movement downward through the lysimeter was modeled with a one-dimensional convective-dispersive solute transport equation:

$$\frac{\partial C}{\partial t} + v \frac{\partial C}{\partial X} = D \frac{\partial^2 C}{\partial X^2} - KC \quad (1)$$

Boundary conditions appropriate to the lysimeter system are:

$$C = C_0 e^{-\gamma t}, \text{ at } X = 0$$

$$C = 0, \text{ at } t = 0 \text{ and } X > 0 \quad (2)$$

$$\frac{\partial C}{\partial X} \rightarrow 0, \text{ as } X \rightarrow +\infty$$

The solution to this differential equation is given by Bear⁴ and Cleary⁵ as:

$$C(X, t) = \frac{C_0 e^{-\gamma t}}{2} \left[\exp \left(\frac{Xv}{2D} - \frac{X}{\sqrt{D}} \sqrt{A} \right) \cdot \operatorname{erfc} \left(\frac{X}{2\sqrt{Dt}} - \sqrt{At} \right) + \exp \left(\frac{Xv}{2D} + \frac{X}{\sqrt{D}} \sqrt{A} \right) \cdot \operatorname{erfc} \left(\frac{X}{2\sqrt{Dt}} + \sqrt{At} \right) \right] \quad (3)$$

Notation used in the above equations is summarized below:

A = Collected terms that account for source diminution:

$$A = \frac{V^2}{4D} + K - \gamma$$

X = Distance in vertical direction (L).

C (X,t) = Concentration of radionuclide at X and t (activity/L³).

C₀ = Initial concentration of radionuclide (activity/L³).

D = Dispersion coefficient (L²/t)

V = Velocity of radionuclide in groundwater (L/t). This is related to the groundwater velocity by

$$V = V_w \left(\frac{1}{1 + K_d R} \right)$$

V_w = Groundwater velocity (L/t)

K_d = Distribution coefficient of soil for radionuclide (L³/mass).

R = Retardation factor = [ρ/η] (Mass/L³)

ρ = Soil bulk density (mass/L³). Assumed = 1.5 g/cm³

η = Effective porosity (dimensionless). Assumed = 0.25

t = Time

K = Radioactive decay constant for radionuclide (1/t)

γ = Source depletion constant (may be a combination of leaching, radioactive decay, etc.) (1/t).

This transport equation is strictly applicable only to situations in which the groundwater velocity is constant. A constant flow velocity would be found only under very simplified saturated flow conditions, which are approached in the saturated lysimeters. In unsaturated flow the hydraulic conductivity (and therefore the flow velocity) is a function of the soil moisture content, which varies in time and space. Therefore, for the equation to be applied to the unsaturated lysimeters, the simplifying assumption is made that the soil moisture content (and therefore the flow velocity) remains constant. Use of the one-dimensional transport

equation for modeling soil columns is well established.^{6,7} The results given in this report represent the first efforts to model lysimeter performance.

Model Predictions

Figures 13 through 18 illustrate the effect of the distribution coefficient (K_d) on the movement of six isotopes through a lysimeter. K_d is a measure of the ability of the soil to impede migration of radionuclides by adsorption or ion exchange. K_d 's vary according to the fraction of sand, silt, or clay present in the soil; the radionuclide; and the physical and chemical properties of the soil and water. K_d 's used for calculation and shown on each figure represent the highest and lowest values expected. Laboratory tests are planned to better define actual K_d 's for SRP soil. Figures 13 through 18 also illustrate the effect of radioactive decay. For example, Figure 13 indicates that for $K_d = 500$, ^{60}Co activity could be detected in soil water about half a foot below the source after 2 years. The "Detection Limit" line on the graphs is based on assumptions of leaching and rainfall as given in Table 2 and discussed later. After 10 years measurable activity would have moved to about one foot below the source. However, after 100 years ^{60}Co would no longer be measurable at any location.

For Figures 13 through 18, a vertical water velocity of 14 ft/yr was assumed. This is about twice the vertical velocity of water in the unsaturated zone of the SRP burial ground, where there is opportunity for run-off to streams.⁸ The lysimeter design does not permit run-off. This water is assumed to percolate downward. Figure 19 shows the effect of these two water velocities on tritium migration. Although no tritium is buried in the large lysimeters, it is an ideal isotope to illustrate flow effects since it has a very small K_d on soil. This figure shows how the concentration at the bottom of the lysimeter varies with time. It also shows the effect of the dispersion coefficient (D). D is primarily dependent on soil particle size, shape, and uniformity. Because the value of D is not known, the relation $D = V\alpha$ was used, where V is the radionuclide velocity, and α is the dispersivity. The value for dispersivity $\alpha = 1$ was used, based on estimates by Gupta and Greenkorn.⁵ Figure 19 indicates C/C_0 is equal to about 0.25 at the lysimeter bottom (3 ft) after only about 0.1 year with a water velocity of 14 ft/yr. Without dispersion, it would require 0.21 years (3 ft ÷ 14 ft/yr) for tritium to reach this level.

Figures 13 through 20 show concentration ratios, but absolute values can be obtained by determining, or assuming, a C_0 . C_0 , the concentration of a radionuclide in the soil water just below the waste, depends on the leach rate and the amount of rainfall.

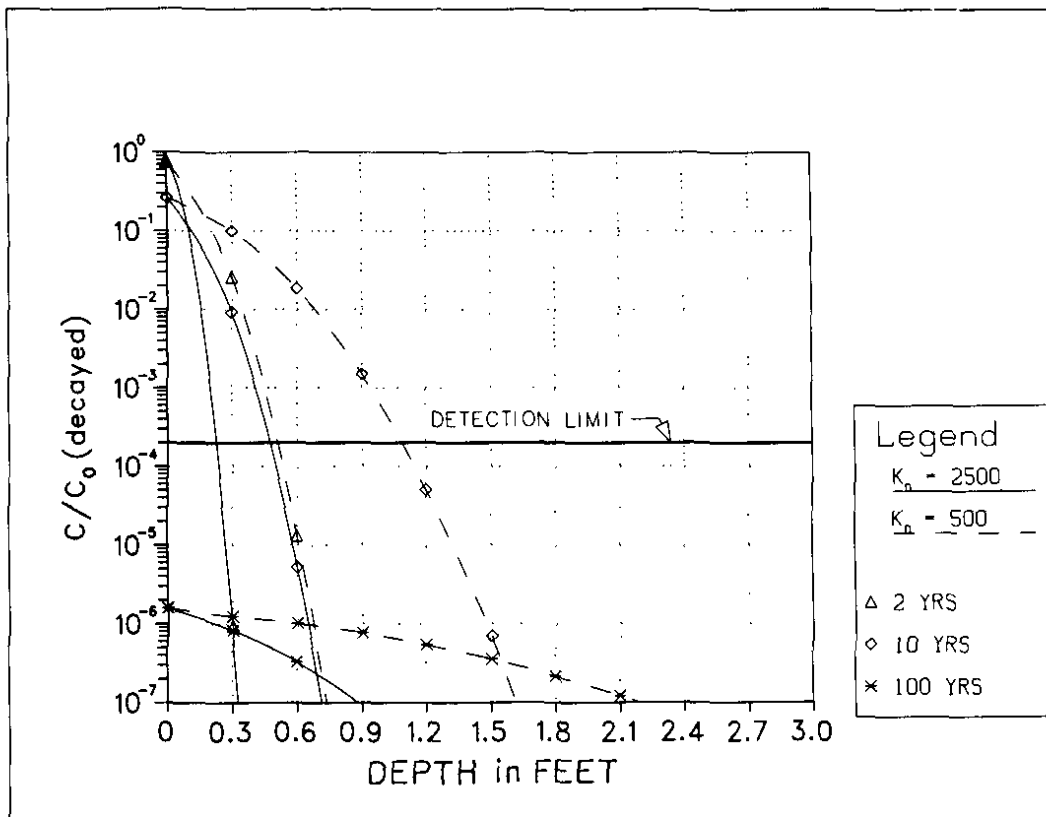


FIGURE 13. Cobalt-60 in Soil Water vs. Depth in Soil

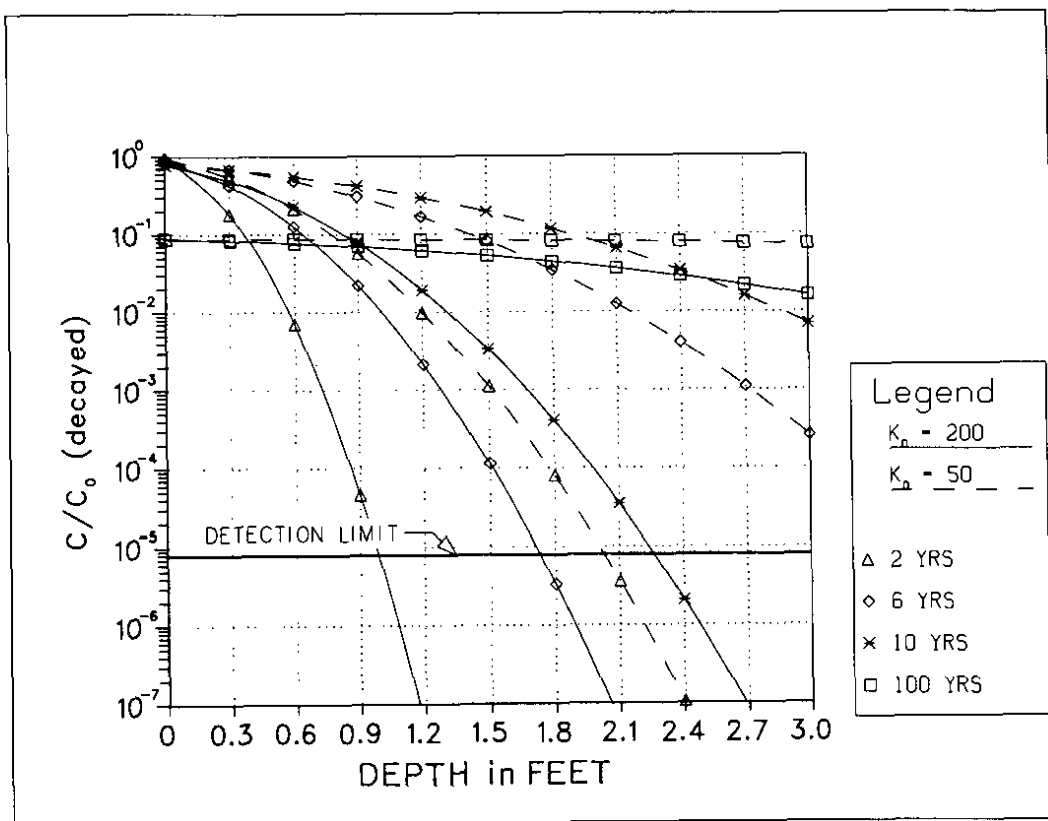


FIGURE 14. Strontium-90 in Soil Water vs. Depth in Soil

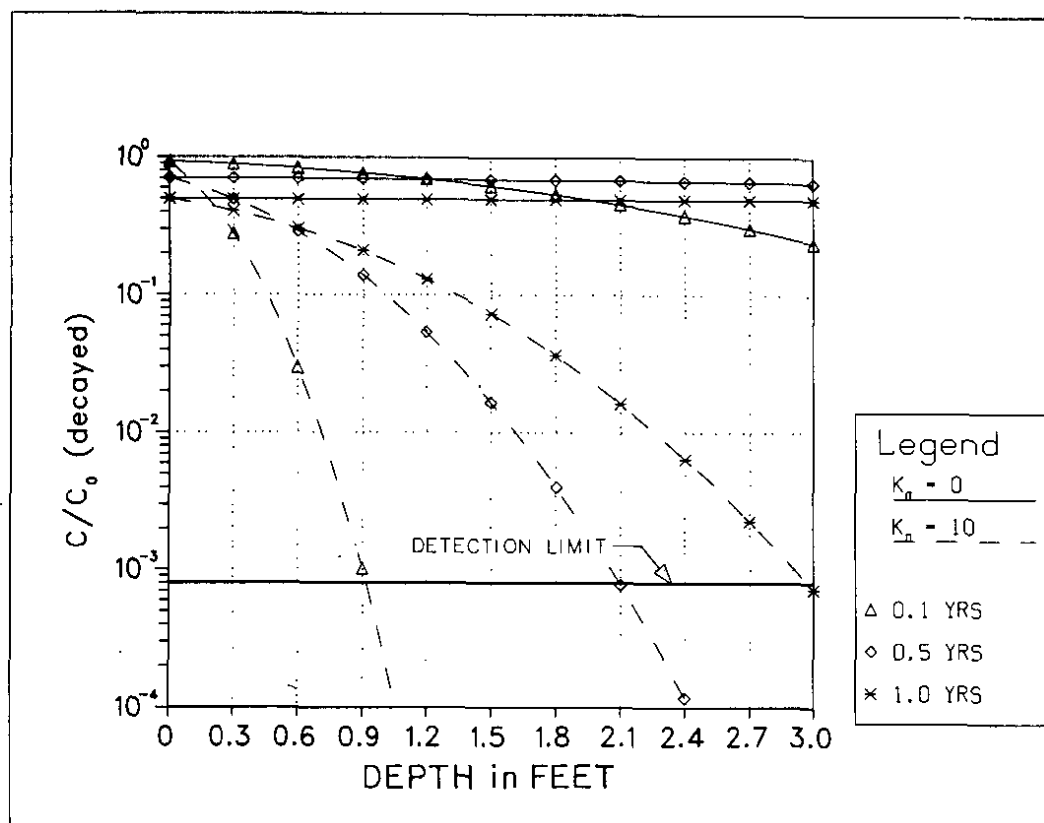


FIGURE 15. Ruthenium-106 in Soil Water vs. Depth in Soil

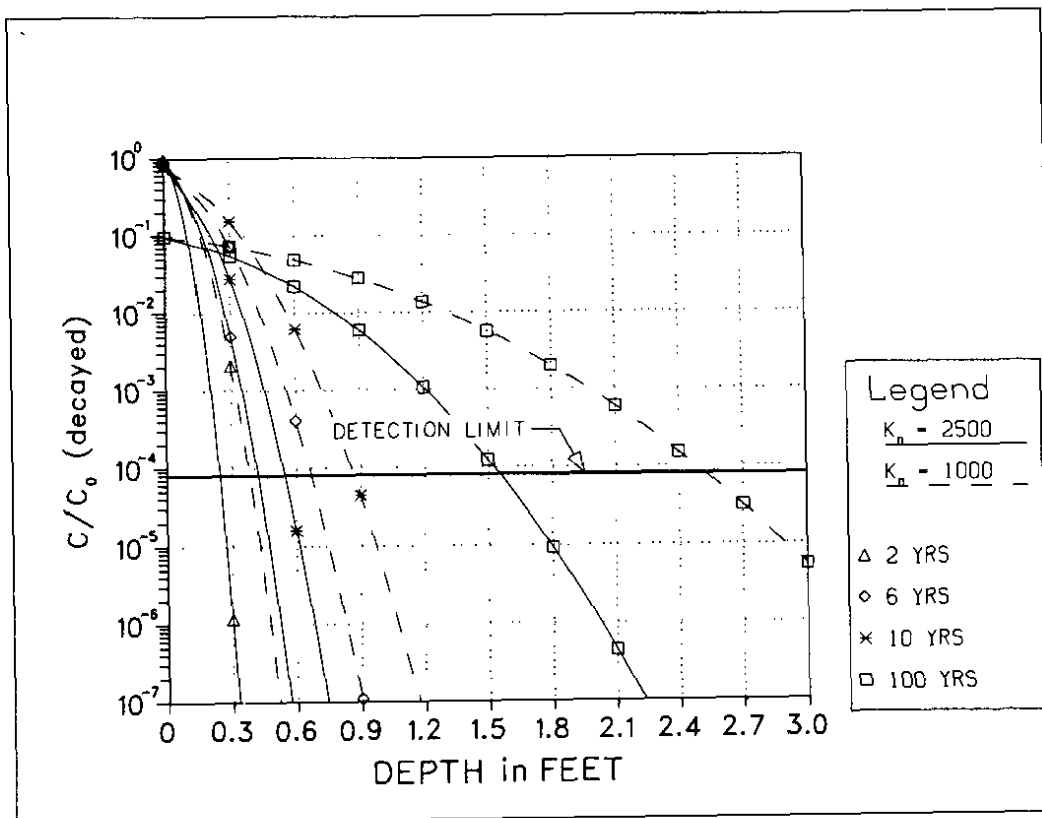


FIGURE 16. Cesium-137 in Soil Water vs. Depth in Soil

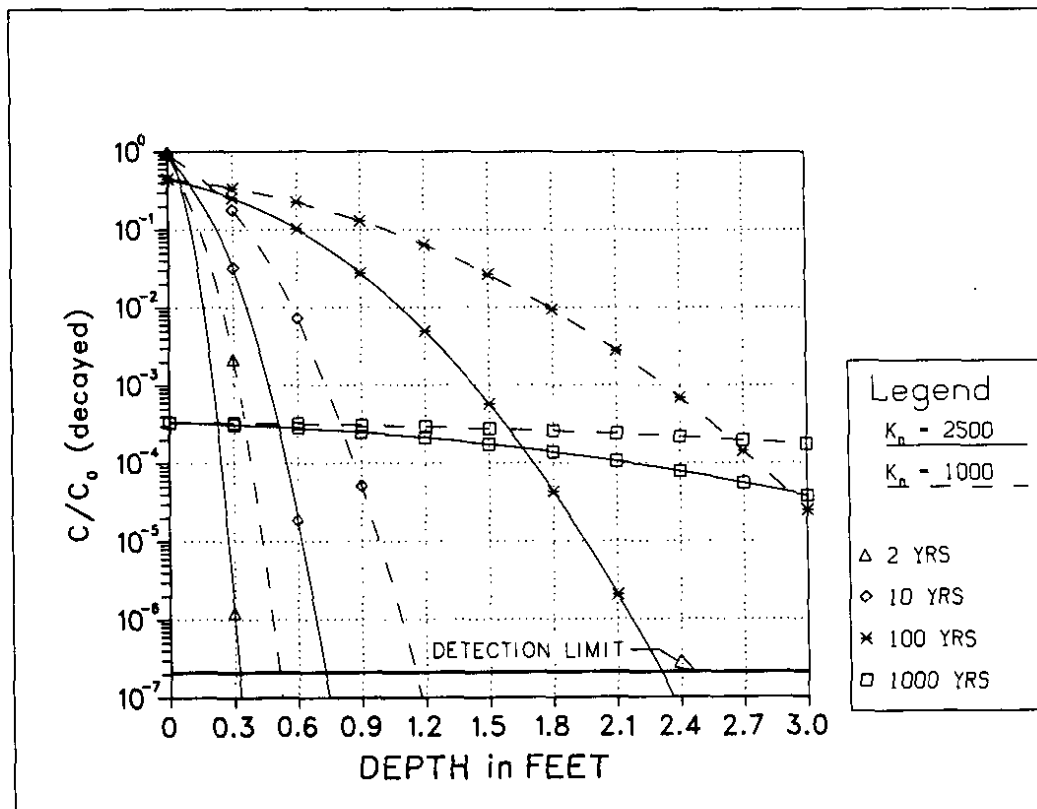


FIGURE 17. Plutonium-238 in Soil Water vs. Depth in Soil

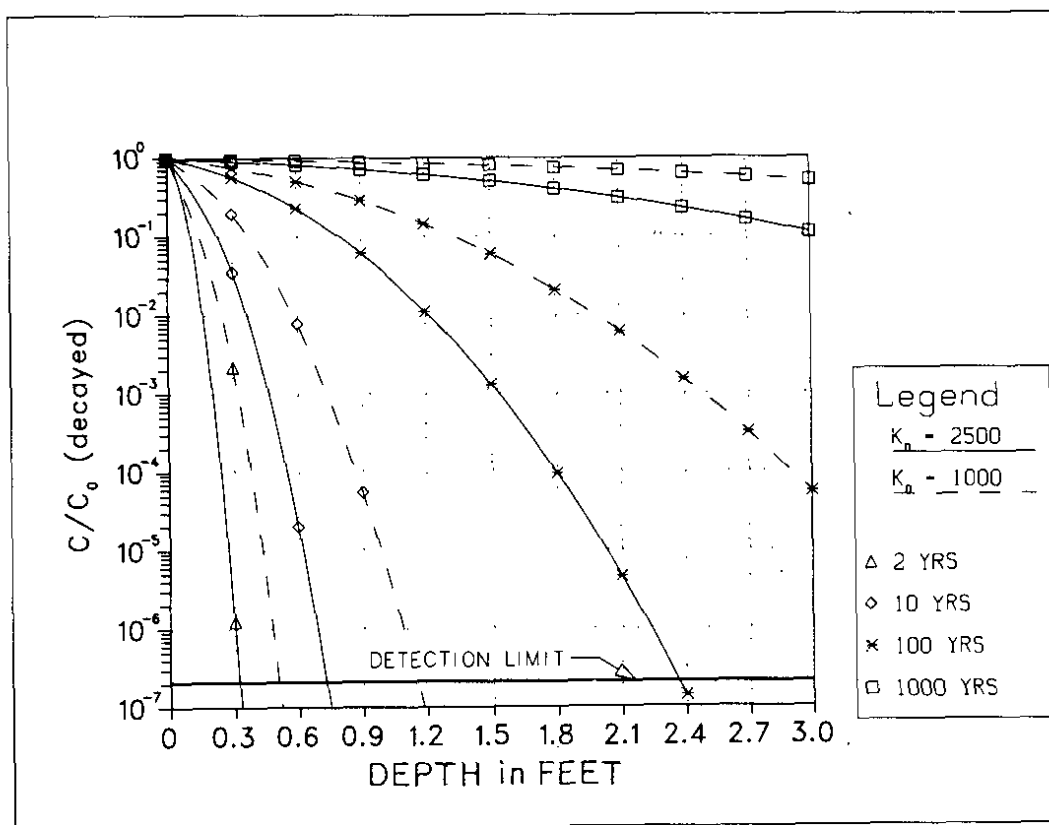


FIGURE 18. Plutonium-239 in Soil Water vs. Depth in Soil

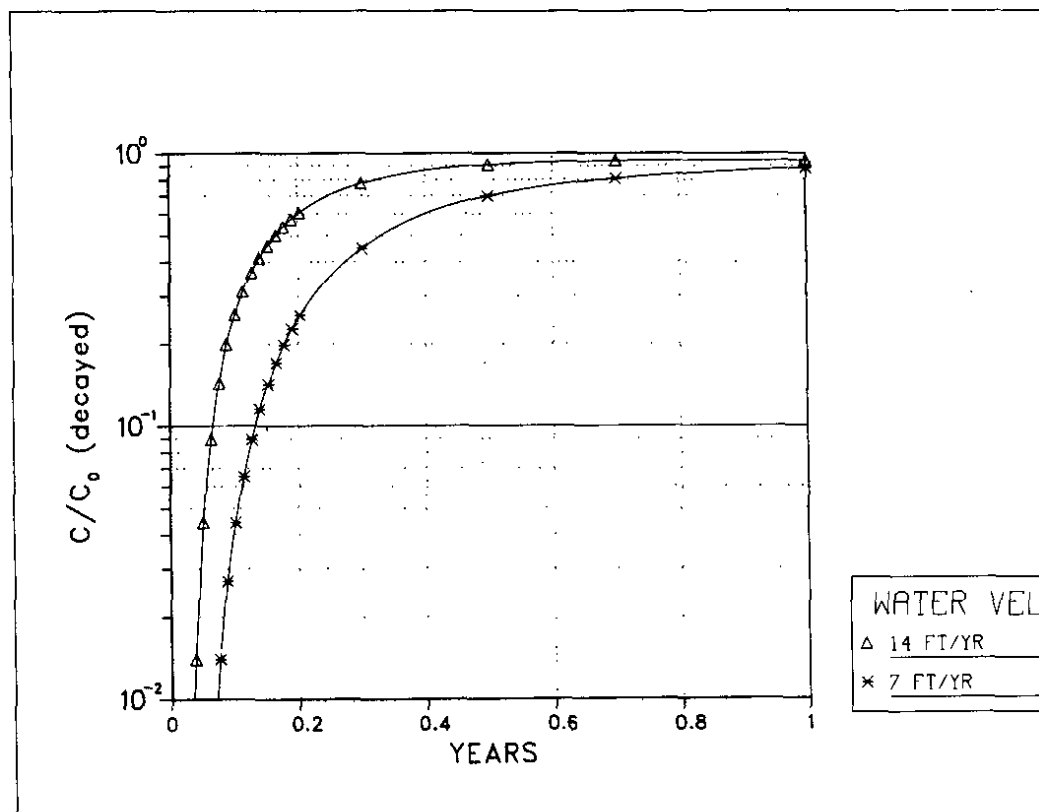


FIGURE 19. Tritium in Soil Water vs. Time
(at bottom of Lysimeter)

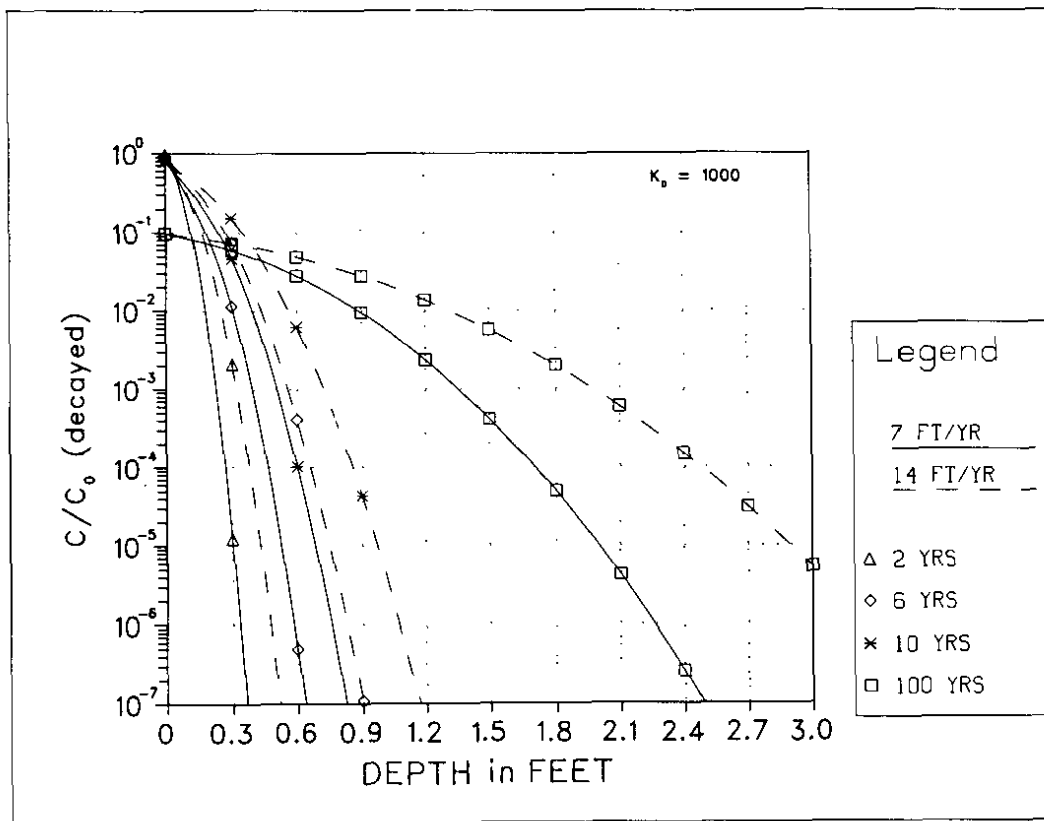


FIGURE 20. Cesium-137 in Soil Water vs. Depth in Soil
(effect of Water Velocity)

TABLE 2

Estimated Minimum C/C_0 for Detection of Radioisotopes

| Radioisotope | Ci per Lysimeter (basis for calculation) | Assumed Fraction Leached, %/yr | Calculated C_0 , pCi/mL ^a | Detection Limited, pCi/mL ^b | C/C_0 (min detectable) ^c |
|---------------------------|---|---|---|--|--|
| $^{238}, ^{239}\text{Pu}$ | 0.2 | 1 | 950 | 2×10^{-4} | 2×10^{-7} |
| ^{137}Cs | 0.1 | 5 | 2380 | 0.2 | 8×10^{-5} |
| ^{90}Sr | 0.1 | 5 | 2380 | 0.02 | 8×10^{-6} |
| ^{60}Co | 0.2 | 1 | 950 | 0.2 | 2×10^{-4} |
| ^{106}Ru | 0.1 | 5 | 2380 | 2 | 8×10^{-4} |

a Basis: 31 in./yr rainfall infiltration.

b Sample size = 500 mL.

c For 6-ft diameter lysimeters. For 10-ft diameter lysimeters multiply listed C/C_0 values by 3.

Since the rainfall is known, if one assumes a leach rate, a value for C_0 can be calculated. Table 2 gives C_0 values based on assumed leach rates and also provides the minimum concentration of the given radionuclide which can be readily detected in water samples. Based on the rainfall and assumed leach rates, the C/C_0 values corresponding to this minimum detectable concentration were determined and are shown as a line on the plots of Figures 13 through 18. Theoretically, any C/C_0 value above this line can be readily detected by standard analytical techniques.

Migration of ^{137}Cs at water velocities of 7 ft/yr and 14 ft/yr were calculated (Figure 20). The velocity of 7 ft/yr approximates the infiltration rate of water in the unsaturated zone in the burial ground (with run-off) and also approximates the saturated zone flow rate in the lysimeters. The effect of water velocity is observed by noting that after 100 years the C/C_0 ratio is 10^{-4} at the 1.7-ft depth with a V_w of 7 ft/yr, but this same concentration is reached at a depth of 2.5 ft after 100 years with a V_w of 14 ft/yr. This comparison also represents the approximate difference in migration between the unsaturated (14 ft/yr) and saturated zones (7 ft/yr) of the lysimeters.

The influence of leaching was neglected in the foregoing plots, as mentioned earlier. However, the transport equation provides for depletion of the source by leaching. (The term γ is the decay constant for the source, considering all mechanisms.) To investigate the effect of depletion of the source by leaching, the distribution of ^{239}Pu in soil water was calculated assuming 1% per year leaching (Figure 21). The case for no depletion of the source by leaching is also shown. The initial concentration (at time = 0) was assumed to be the same for both cases. Concentrations are presented in absolute units based on the 1% leach rate and the average rainfall (Table 2). The results show that leaching and decay reduces the source concentration by a factor of about 0.33 after 100 years, but by a much smaller factor (3.6×10^{-5}) after 1000 years. However, because of the slow rate of movement of ^{239}Pu through soil, the effect is only a factor of about 4 at the 3-ft depth after 1000 years.

Recommendations for Field Tests

This study suggests that data from the lysimeters may be used in the following way to determine important parameters of migration:

- D - The dispersion coefficient for local soils could be measured by introducing a nonadsorbing species (e.g. HTO or T_2O) to a new (or existing) lysimeter and observing the effluent concentration. With $K_d = 0$ and C_0 known, D could be determined by substituting values into the equation to determine which value best fits the data.

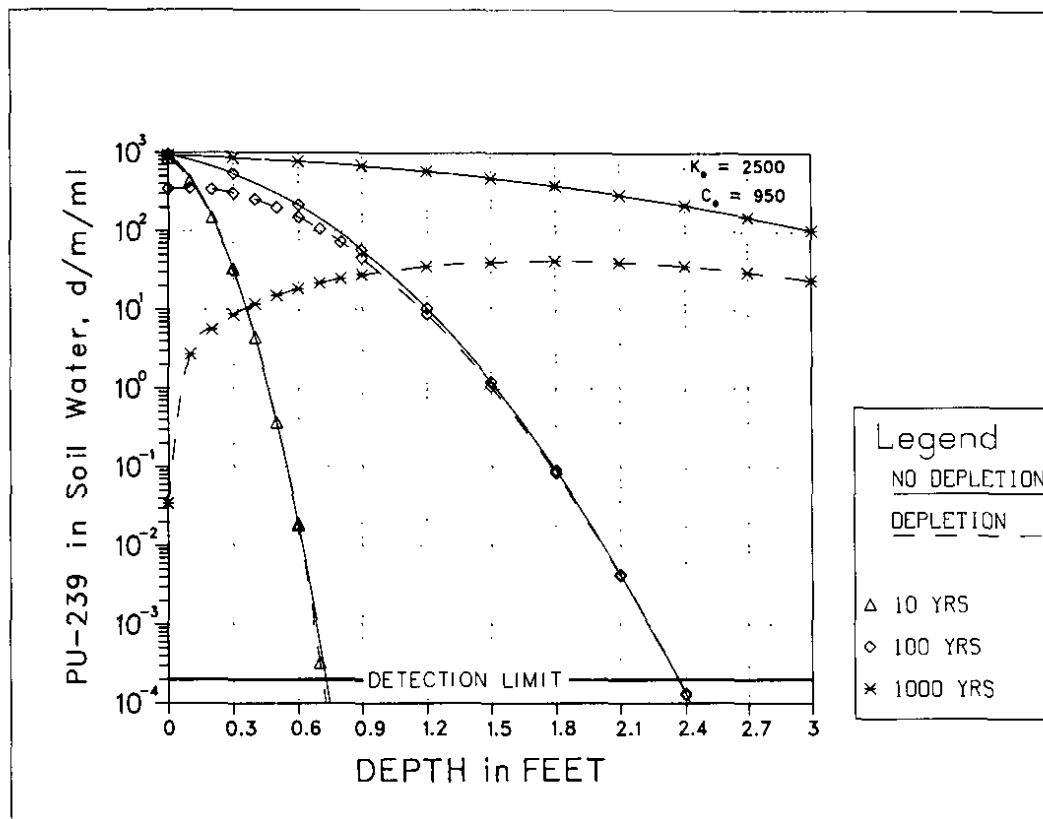


FIGURE 21. Plutonium-239 in Soil Water vs. Depth in Soil (effect of source depletion)

- K_d - K_d values could be determined (assuming D is known) by substituting various K_d values into the transport equation until the best agreement was found between calculated and measured lysimeter effluent data. Figure 22 shows calculated effluent data for ^{106}Ru plotted vs. time, holding D constant. The slope and shape of the curves show a strong dependence on K_d . For isotopes which are not detectable in the lysimeter sump water, soil moisture samples taken at different times at a single depth would provide similar information. One could also sample soil moisture at various depths at a given time and insert the radionuclide concentrations into the transport equation and solve for K_d .
- C_0 - (or leaching) - C_0 becomes known when one establishes the C/C_0 vs. depth relationship (by determining K_d and D) and also obtains an actual concentration value for a specific case.

The foregoing discussion illustrates the type of calculations which can be made and identifies important parameters. The calculations presented are based on ranges of some parameters and estimates of others. However, the information given should provide guidance for directing further studies and for anticipating lysimeter performance until improved input values are obtained.

CONCLUSIONS

The assessment of potential long-term hazards from buried radioactive wastes is important in making good decisions about the future disposition of burial grounds. The lysimeter studies described in this report are designed to provide improved radionuclide migration data for these assessments. Results from the studies will also be useful in guiding ongoing burial ground operations. The work should be carried out as planned and perhaps amplified as results become available and as future plant operations change.

The need for additional work to define some of the parameters which influence radionuclide migration is emphasized. This work should consider the unsaturated and the saturated zone. Soil used in any experiments should be well characterized. Parameters which need further study are:

- K_d (distribution coefficient) - needed for various radionuclides and various soil types.

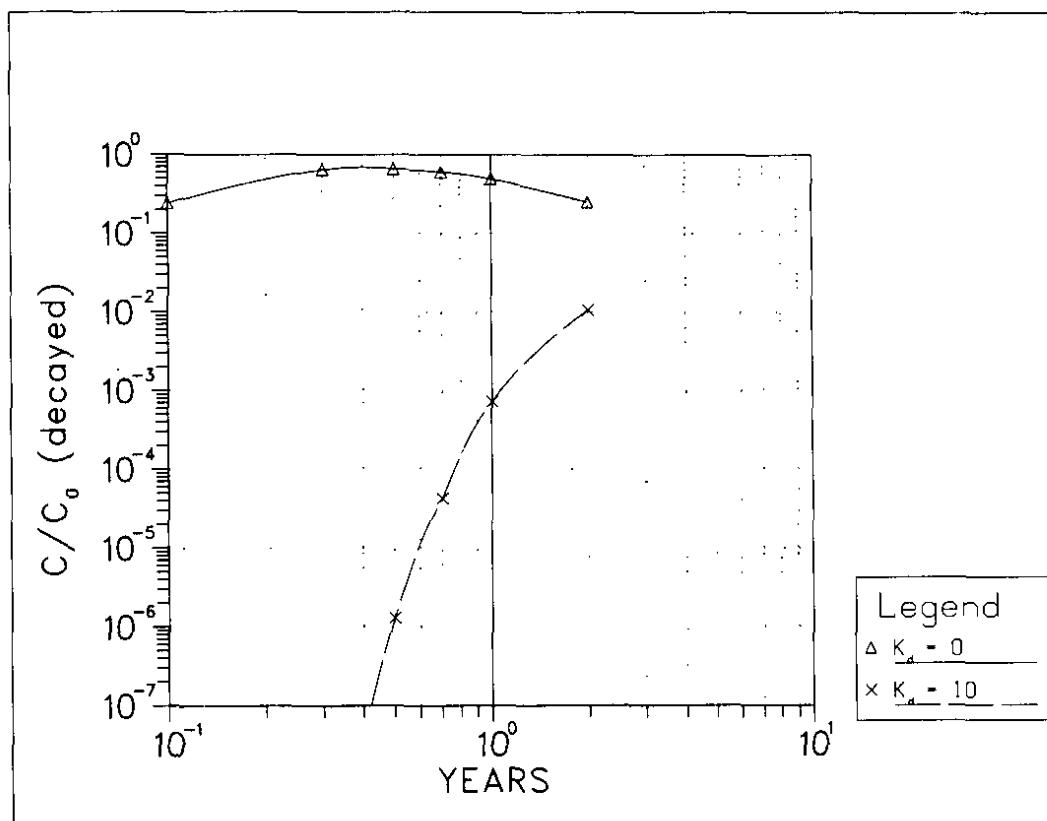


FIGURE 22. Ruthenium-106 in Sump Water
(effect of K_d)

- D (dispersion coefficient) - this parameter has a strong influence on migration, yet very little is known about it for local soil types.
- V_w (groundwater velocity) - including the effect of the intermittent nature of rainfall on radionuclide migration.

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APPENDIX

TABLE A-1

Comprehensive Lysimeter Loading Data

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | | | | |
|--|------------------|---------------------|--|---|-------|--------------------------------------|-------------------|---------|
| Separations Laboratory Glove Box Waste | 1 | June 1978 | Cotton pads, plastic bags, paper, rubber gloves, glass & plastic bottles (tops removed). Two 5-gallon cardboard ice cream cartons of waste placed in each lysimeter. | ²³⁸ Pu | 0.45 | | | |
| | 2 | June 1978 | Same as Lysimeter 1 | ²³⁸ Pu | 0.48 | | | |
| | 3 | June 1978 | Same as Lysimeter 1 | ²³⁸ Pu | 0.47 | | | |
| | 4 | June 1978 | Same as Lysimeter 1 | ²³⁸ Pu | 0.49 | | | |
| | 5 | June 1978 | Same as Lysimeter 1 | ²³⁸ Pu | 0.53 | | | |
| | | | | | | | | |
| Reactor Moderator Deionizer Resin | 6 | July 1978 | 1 ft ³ of resin used for removing ionic contaminants from D ₂ O moderator. Resin is 5 vol parts anion (Rohm & Haas "Amberlite" IRA 400C) and 1 vol part cation (Rohm & Haas "Amberlite" IR 120) resin. | Average Analyses, ² Ci | | Individual Resin Sample Analyses, Ci | | |
| | | | | ¹⁴ C | 0.35 | [B] | ⁶⁰ Co | 0.18 |
| | | | | ⁵⁴ Mn | 0.005 | | ⁹⁰ Sr | 0.012 |
| | | | | ⁶⁰ Co | 0.27 | | ¹³⁷ Cs | 0.018 |
| | | | | ⁶⁵ Zn | 0.008 | [C] | ¹⁴⁴ Ce | 0.013 |
| | | | | ¹³⁴ Cs | 0.002 | | Pu(ext) | 0.00028 |
| | | | | ¹³⁷ Cs | 0.023 | | Gross α | 0.00029 |
| | | | | ¹⁴⁴ Ce | 0.86 | | | |
| | | | | | | | | |

TABLE A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ | |
|-----------------------------------|------------------|---------------------|---------------------|------------------------------------|---|
| | | | | Average Analyses, ² Ci | Individual Resin Sample Analyses, Ci |
| Reactor Moderator Deionizer Resin | 7 | July 1978 | Same as Lysimeter 6 | Same as Lysimeter 6 | ⁶⁰ Co 0.17 ⁹⁰ Sr 0.011 ¹³⁷ Cs 0.016 ¹⁴⁴ Ce 0.012 Pu(ext) 0.00014 Gross α 0.00029 |
| | | | | | [F] |
| | 8 | July 1978 | Same as Lysimeter 6 | Same as Lysimeter 6 | ⁶⁰ Co 0.12 ⁹⁰ Sr 0.0069 ¹³⁷ Cs 0.016 ¹⁴⁴ Ce 0.011 Pu(ext) 0.00036 Gross α 0.00037 |
| | | | | | [F] |
| | 9 | July 1978 | Same as Lysimeter 6 | Same as Lysimeter 6 | ⁶⁰ Co 0.086 ⁹⁰ Sr 0.0063 ¹³⁷ Cs 0.011 ¹⁴⁴ Ce 0.0076 Pu(ext) 0.00020 Gross α 0.00028 |
| | | | | | [F] |
| | 10 | July 1978 | Same as Lysimeter 6 | Same as Lysimeter 6 | ⁵⁴ Mn 0.0011 ⁵⁷ Co 0.00061 ⁶⁰ Co 0.10 ⁹⁰ Sr 0.0031 ¹³⁷ Cs 0.015 ¹⁴⁴ Ce 0.0052 Pu(ext) 0.000083 Gross α 0.00015 |
| | | | | | [F] |

TABLE A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|---|------------------|---------------------|--|---|---------|
| 239Pu Finishing Line (JB-Line) Job Control Waste | 11 | April 1979 | Plastic bags, Pu oxide, filter | 239Pu | 0.36 |
| | 12 | April 1979 | Plastic bags, Pu oxide | 239Pu | 0.34 |
| | 13 | April 1979 | Plastic bags, gaskets, Pu oxide, gloves | 239Pu | 0.36 |
| | 14 | April 1979 | Plastic bags, Pu oxide, vacuum cleaner, gasket | 239Pu | 0.37 |
| | 15 | April 1979 | Plastic bags, Pu oxide, tools | 239Pu | 0.36 |
| NOTE: In three 5-gallon metal paint buckets - waste removed from buckets. | | | | | |
| SRL High Level Caves (HLC) and Actinide Materials Facility (AMF) Waste | 16 | January 1980 | AMF Waste - Cotton pads, paper, plastic, cloth gloves, tools. | 238Pu | 0.16 |
| | | | | 241Am | 0.00002 |
| | | | HLC Waste - Plastic bags, cotton pads, plastic bottles, grinding paper, dust stop filter, manipulator fingers. | 60Co | 0.0021 |
| | | | | 106Ru | 0.0035 |
| | | | | 125Sb | 0.0012 |
| | | | | 134Cs | 0.0035 |
| | | | | 137Cs | 0.023 |

[A]

[D]

[C]

TABLE A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|---|------------------|---------------------|--|---|-------|
| SRL High Level Caves (HLC) and Actinide Materials Facility (AMF) Waste (Cont'd) | 17 | February 1980 | AMF Waste - Same as Lysimeter 16 | ²³⁸ Pu 0.16 | } [D] |
| | | | | ²⁴¹ Am 0.0007 Ci | |
| | 18 | January 1980 | HLC Waste-Assorted plastic bags, cotton pads, sample vials and caps, metal knife, plastic bottles, Zr powder | ¹⁰⁶ Ru 0.026 | } [C] |
| | | | | ¹³⁴ Cs 0.027 | |
| | | | | ¹³⁷ Cs 0.163 | |
| | | | AMF Waste-Plastic bags, cotton pads | ²³⁸ Pu 0.14 | } [D] |
| | | | | ²⁴¹ Am 0.000015 | |
| | 19 | January 1980 | HLC Waste-Plastic bags, cotton pads, used centrifuge cover, broken glass bottles, polyethylene bottles, broken manipulator hand. | ²² Na 0.00005 | } [C] |
| | | | | ⁶⁰ Co 0.00022 | |
| | | | | ¹⁰³ Ru 0.078 | |
| | | | | ¹⁰⁶ Ru 0.037 | |
| | | | | ¹²⁵ Sb 0.0043 | |
| | | | | ¹³⁴ Cs 0.0063 | |
| | | | | ¹³⁷ Cs 0.037 | |
| | | | | ¹⁵⁴ Eu 0.0016 | |
| | | | AMF Waste-plastic bags, cotton pads, glassware | ²³⁸ Pu 0.16 | } [D] |
| | | | | ²⁴¹ Am 0.000032 | |
| | | | HLC Waste-Dust stop filter, plastic bags, cotton wipes, polishing cloths | ⁶⁰ Co 0.0010 | } [C] |
| | | | | ¹⁰⁶ Ru 0.00060 | |
| | | | | ¹³⁴ Cs 0.00026 | |
| | | | | ¹³⁷ Cs 0.00165 | |
| | | | | ¹⁴⁴ Ce 0.00090 | |

TABLE A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|---|------------------|---------------------|---|---|-------|
| SRL High Level Caves (HLC) and Actinide Materials Facility (AMF) Waste (Cont'd) | 20 | January 1980 | AMF Waste-Plastic bags, cotton pads. | ²³⁸ Pu 0.30 | } [D] |
| | | | | ²⁴¹ Am 0.00016 | |
| | | January 1980 | HLC Waste-Plastic bags, cotton pads, glass beakers, centrifuge cover, miscellaneous glassware & polyethylene bottles. | ²² Na 0.00097 | } [C] |
| | | | | ⁶⁰ Co 0.00047 | |
| | | | | ⁹⁵ Zr 0.0011 | |
| | | | | ¹⁰³ Ru 0.095 | |
| | | | | ¹⁰⁶ Ru 0.0021 | |
| | | | | ¹³⁴ Cs 0.00061 | |
| | | | | ¹³⁷ Cs 0.041 Ci | |
| | | | | ¹⁵⁴ Eu 0.0032 Ci | |
| Canyon Pipe Jumpers | 21 | December 1979 | Pipe jumper pieces from H-Area warm canyon. Jumpers had been exposed to Pu and Np solutions. Two lengths of 2-in pipe x 2 1/2 ft. long and three lengths of 3-in pipe x 2 1/2 ft. long | ²³⁸ Pu 0.0074 | } [E] |
| | | | | ²³⁹ Pu 0.00093 | |
| | | | | ⁹⁰ Sr 2.8×10^{-6} | |
| | | | | ⁹⁵ Zr 2.5×10^{-6} | |
| | | | | ⁹⁵ Nb 6.4×10^{-6} | |
| | | | | ¹³⁷ Cs 1.8×10^{-6} | |
| | 22,23,24,25 | December 1979 | Each same as Lysimeter 21 | Each same as Lysimeter 21 | |
| Separations Laboratory Waste | 26 | August 1979 | Glass sample vials, vial caps, burette and pipette tips, cotton pads and plastic containers. One 5-gallon cardboard ice cream carton containing three 2 quart cardboard cartons per lysimeter. All containers were opened in lysimeter. | ⁹⁵ Zr 0.0077 | } [C] |
| | | | | ⁹⁵ Nb 0.011 | |
| | | | | ¹⁰⁶ Ru 0.012 | |
| | | | | ¹³⁴ Cs 0.0017 | |
| | | | | ¹³⁷ Cs 0.0054 | |
| | | | | ¹⁴⁴ Ce 0.050 | |

TABLE A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|------------------------------|------------------|---------------------|----------------------|---|---------|
| Separations Laboratory Waste | 27 | August 1979 | Same as Lysimeter 26 | ⁹⁵ Zr | 0.018 |
| | | | | ⁹⁵ Nb | 0.057 |
| | | | | ¹⁰⁶ Ru | 0.015 |
| | | | | ¹²⁵ Sb | 0.0028 |
| | | | | ¹³⁴ Cs | 0.0013 |
| | | | | ¹³⁷ Cs | 0.0049 |
| | | | | ¹⁴⁴ Ce | 0.040 |
| | | | | [C] | |
| | 28 | August 1979 | Same as Lysimeter 26 | ⁹⁵ Zr | 0.021 |
| | | | | ⁹⁵ Nb | 0.022 |
| | | | | ¹⁰⁶ Ru | 0.0020 |
| | | | | ¹³⁴ Cs | 0.0028 |
| | | | | ¹³⁷ Cs | 0.0085 |
| | | | | ¹⁴⁴ Ce | 0.078 |
| | | | | [C] | |
| | 29 | August 1979 | Same as Lysimeter 26 | ⁹⁵ Zr | 0.0060 |
| | | | | ¹⁰³ Ru | 0.00091 |
| | | | | ¹⁰⁶ Ru | 0.014 |
| | | | | ¹³⁴ Cs | 0.00099 |
| | | | | ¹³⁷ Cs | 0.0043 |
| | | | | ¹⁴⁴ Ce | 0.040 |
| | | | | [C] | |
| | 30 | August 1979 | Same as Lysimeter 26 | ⁹⁵ Zr | 0.0052 |
| | | | | ⁹⁵ Nb | 0.0084 |
| | | | | ¹⁰⁶ Ru | 0.0087 |
| | | | | ¹³⁴ Cs | 0.0011 |
| | | | | ¹³⁷ Cs | 0.0037 |
| | | | | ¹⁴⁴ Ce | 0.035 |
| | | | | [C] | |

Table A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|--------------------------------------|------------------|---------------------|--|---|-------|
| Separations Canyon Job Control Waste | 31 | May 1979 | Plastic suits, cotton coveralls, rubber gloves and over shoes & cotton pads. Three 15"x15"x15" cardboard boxes of waste were placed in each lysimeter. Boxes were opened in lysimeter. | ⁹⁵ Zr | 0.11 |
| | | | | ⁹⁵ Nb | 0.17 |
| | | | | ¹⁰⁶ Ru | 0.09 |
| | | | | ¹³⁷ Cs | 0.16 |
| | | | | ¹⁴⁴ Ce | 0.38 |
| | | | | } | |
| | | | | [C] | |
| | 32 | May 1979 | Same as Lysimeter 31 | ⁹⁵ Zr | 0.28 |
| | | | | ⁹⁵ Nb | 0.27 |
| | | | | ¹⁰³ Ru | 0.24 |
| | | | | ¹⁰⁶ Ru | 0.29 |
| | | | | ¹³⁷ Cs | 0.21 |
| | | | | ¹⁴⁴ Ce | 0.65 |
| | | | | } | |
| | | | | [C] | |
| | 33 | May 1979 | Same as Lysimeter 31 | ⁹⁵ Zr | 0.16 |
| | | | | ⁹⁵ Nb | 0.21 |
| | | | | ¹⁰⁶ Ru | 0.28 |
| | | | | ¹³⁷ Cs | 0.15 |
| | | | | ¹⁴⁴ Ce | 0.29 |
| | | | | } | |
| | | | | [C] | |
| | 34 | May 1979 | Same as Lysimeter 31 | ⁹⁵ Zr | 0.10 |
| | | | | ⁹⁵ Nb | 0.13 |
| | | | | ¹⁰⁶ Ru | 0.19 |
| | | | | ¹³⁷ Cs | 0.16 |
| | | | | ¹⁴⁴ Ce | 0.28 |
| | | | | } | |
| | | | | [C] | |
| | 35 | May 1979 | Same as Lysimeter 31 | ⁹⁵ Zr | 0.09 |
| | | | | ⁹⁵ Nb | 0.14 |
| | | | | ¹⁰⁶ Ru | 0.15 |
| | | | | ¹²⁵ Sb | 0.041 |
| | | | | ¹³⁷ Cs | 0.14 |
| | | | | ¹⁴⁴ Ce | 0.37 |
| | | | | } | |
| | | | | [C] | |

Table A-1, Contd

| Waste Form | Lysimeter Number | Date Waste Emplaced | Waste Description | Radioactivity Content ¹ , Ci | |
|----------------------------|------------------|---------------------|---|---|-------|
| Reactor Scrap Metal | 36 | March 1978 | Reactor scrap metal ³ consisting of cut-up aluminum fuel housings and stainless steel snap rings | ⁴⁶ Sc 0.0074 ⁵⁴ Mn 0.036 ⁶⁰ Co 0.22 ⁶⁵ Zn 0.71 | } [C] |
| | 37 | March 1978 | Same as Lysimeter 36 | ⁴⁶ Sc 0.0062 ⁵⁴ Mn 0.031 ⁶⁰ Co 0.17 ⁶⁵ Zn 0.50 | } [C] |
| | 38 | March 1978 | Same as Lysimeter 36 | ⁴⁶ Sc 0.009 ⁵⁴ Mn 0.052 ⁶⁰ Co 0.19 ⁶⁵ Zn 0.56 | } [C] |
| | 39 | March 1978 | Sames as Lysimeter 36 | ⁴⁶ Sc 0.0124 ⁵⁴ Mn 0.064 ⁶⁰ Co 0.20 ⁶⁵ Zn 0.62 | } [C] |
| | 40 | June 1979 | Same as Lysimeter 36 | ⁵⁴ Mn 0.0095 ⁶⁰ Co 0.15 ⁶⁵ Zn 0.17 | } [C] |
| Control Lysimeters C1 & C2 | | None | None | None | |

See next page for footnotes.

Footnotes for Table A-1

1. The letters in brackets in the table indicate the technique (listed below) used for measuring the radioactivity content of wastes emplaced in the lysimeters.
 - [A] = Counted with the Separations Control Laboratory (Bldg. 772-F) Waste Monitor, employing a NS-633 multi-channel analyzer and an ORTEC 8108-0320 Ge-Li detector. ^{239}Pu was detected by counting the 129 mev photons, and ^{238}Pu was detected by counting the 153 mev photons.
 - [B] = ^{14}C was extracted from moderator resin by bicarbonate exchange and counted by liquid scintillation. Finally, combustion of the resin and collection and counting of CO_2 showed no residual ^{14}C in the resin.
 - [C] = Counted with the gamma monitor used for surveying incoming waste at the burial ground. This monitor has been evaluated and described.
 - [D] = Counted with the Savannah River Laboratory Waste Monitor, employing a Canberra 8100-4K multi-channel analyzer, a Canberra 7705 planar LC detector and a PDP-11-16A computer processing unit. The 17 kev photons were used for detecting ^{238}Pu and the 59.5 kev photons were used for detecting ^{241}Am .
 - [E] = Three-inch long samples were taken from both ends and the middle of both pipe jumpers used. These samples were counted with an alpha probe, then the radioactivity was dissolved from them with three successive exposures to $\text{HNO}_3\text{-HF}$ solution. Probe data and final leach solution data both indicated no significant residual radioactivity on the samples. The curies present on the samples was obtained by analyses of the leach solution. The curies of radioactivity placed in the lysimeters was determined by assuming that the amount of radioactivity on the pipe was uniform and thus directly proportional to the length of the pipe.
 - [F] = Standard laboratory and counting room techniques, possibly including chemical separations, alpha or beta counting, gamma spectrometry, etc.
2. Based on simultaneous analysis of packages for all five lysimeters.
3. Calculations have shown that certain non-gamma emitting radioisotopes are likely to be present [e.g. ^{63}Ni (100 yrs);⁹ and ^{55}Fe (2.7 yrs)] due to neutron capture reactions, but the amount is uncertain due to variations in the composition of aluminum and stainless steels used in reactor components. The amount of these radioisotopes in the scrap loaded into the lysimeters was not determined.

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