SRNL-STI-2012-00299 Revision 0

Decontamination Factors and Filtration Flux Impact to ARP at Reduced MST Concentration

D. T. Hobbs D. T. Herman M. R. Poirier

June 2012

Savannah River National Laboratory Savannah River Nuclear Solutions, LLC Aiken, SC 29808

Prepared for the U.S. Department of Energy under contract number DE-AC09-08SR22470.

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REVIEWS AND APPROVALS

AUTHORS:

EXECUTIVE SUMMARY

Tank Farm and Closure Engineering is evaluating changes to the Actinide Removal Process facility operations to decrease the MST concentration from 0.4 g/L to 0.2 g/L and the contact time from 12 hours to between 6 and 8 hours. For this evaluation, SRNL reviewed previous datasets investigating the performance of MST at 0.2 g/L in salt solutions ranging from 4.5 to 7.5 M in sodium concentration. In general, reducing the MST concentration from 0.4 to 0.2 g/L and increasing the ionic strength from 4.5 to 7.5 M in sodium concentration will decrease the measured decontamination factors for plutonium, neptunium, uranium and strontium. The decontamination factors as well as single standard deviation values for each sorbate are reported. These values are applicable within the sorbate and sodium concentrations used in the experimental measurements.

Decreasing the MST concentration in the ARP from 0.4 g/L to 0.2 g/L will produce an increase in the filter flux, and could lead to longer operating times between filter cleaning. The increase in flux is a function of a number of operating parameters, and is difficult to quantify. However, it is estimated that the reduction in MST could result in a reduction of filtration time of up to 20%.

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1.0 Introduction

Monosodium titanate (MST) is used in the Actinide Removal Process (ARP) facility to remove 90 Sr and alpha-emitting radionuclides, principally the radioisotopes of plutonium, neptunium, and uranium. The current ARP operation adds MST at a concentration of 0.4 g/L to each batch of waste processed in the 241-96H strike tanks. The reaction suspension is mixed for 12 hours and then transferred to the 512-S facility for filtration to concentrate the MST solids. The clarified filtrate from the crossflow filters transfers into the Modular Caustic Side Solvent Extraction Unit (MCU) for removal of cesium.

Tank Farm and Closure Engineering (TF&CE) are evaluating modifications to the ARP flowsheet to increase throughput. These modifications include reducing the MST concentration and reducing the strike time to eight hours or less. To support this evaluation, TF&CE requested that SRNL determine decontamination factors (DF) for Pu, Np, U and Sr at an MST concentration of 0.2 g/L with contact times of 6 and 8 hours in salt solutions having sodium concentrations ranging from 4.5 M to 7.5 M.¹ SRNL was also requested to determine the effect of the reduced MST concentration on the 512-S crossflow filtration rate. This report summarizes the findings of these evaluations.

2.0 Experimental Procedure

No experimental work was performed in support of the Technical Task Request $(TTR)^1$

3.0 Results and Discussion

3.1 Sorbate Removal by MST

The DF for a sorbate is the ratio of the initial solution concentration and the concentration measured at some time after the introduction of the MST. Parameters that influence DF values include the initial sorbate concentration, the ionic strength or sodium ion concentration of the solution, the MST concentration, and the temperature. Temperature is controlled in the ARP process strike tanks at 25 ± 5 °C.² Previous testing has shown that this small temperature range has minimal influence on MST performance.³ Therefore, this evaluation did not consider the effect of temperature.

The ARP initiated radioactive operations in April 2008 and has processed more than 2.8 million gallons of liquid waste through March of 2012. The liquid phase processed through ARP is staged as large, well-characterized batches in Tank 49H. Table 3-1 provides a summary of sodium, strontium and actinide concentrations for the five macrobatches of waste that have been assembled and processed or will be processed in ARP.³⁻⁸

The target sodium concentration for the feed solution to ARP was originally set at 5.6 M. However, the ARP facility has moved to processing waste solutions having sodium concentrations in excess of 6.0 M. Previous testing has shown that MST performance as measured by the decontamination factor decreases with increasing sodium concentration and lower MST concentration.⁹⁻¹²⁶ Thus, the DF values for Sr, Pu, Np and U would be expected to be lower upon processing waste solutions with higher sodium concentrations.

Currently, the MST strike tanks in ARP are charged with MST at a concentration of 0.4 g/L. TF&CE requested SRNL to provide recommended DF values for each of the sorbates upon decreasing the MST concentration by 50% to 0.2 g/L . As shown previously, sorbate removal (i.e., DF value) decreases upon a decrease in the MST concentration. $9-12$

	MB1	MB2	MB3	MB4	MB ₅
Na(M)	5.26	5.55	6.73	6.57	6.25
Total Sr (ug/L)	$< 2.77E + 01$	$< 7.5E + 01$	$<$ 3.49E+01	$<4.61E+01$	$< 2.42E + 01$
Total Pu (ug/L)	$1.08E + 01$	$4.77E + 00$	$9.29E + 00$	$3.57E + 01$	$2.20E + 01$
Np-237 (ug/L)	$2.54E + 01$	$< 5.01E + 01$	$2.94E + 01$	$2.16E + 01$	$4.81E + 00$
Total U (ug/L)	$1.78E + 03$	$5.30E + 03$	$1.82E + 04$	$2.63E + 04$	$1.11E + 04$

Table 3-1. Sodium and Sorbate Concentrations in ARP Macrobatches 1 – 5.

There have been several experimental data sets that have measured the performance of MST at a concentration of 0.2 g/L with salt solutions having sodium concentrations ranging between 4.5 M and 7.5 M.⁹⁻¹³ These tests were carried out at 25 ± 3 °C, which is very similar to the temperature range specified for ARP operations. Table 3-2 provides the initial sorbate and sodium concentrations for each of the test solutions used in these experiments. Within analytical uncertainty, the strontium, plutonium, neptunium, and uranium concentrations in the test solutions are equal to or exceed those in the waste solutions processed to date in ARP. Note that the uranium is the only sorbate in any of the ARP macrobatch feed solutions that is at or close to the maximum soluble concentration in the simulant and tank waste solutions tested in the laboratory for MST performance. Thus, the results reported in this document are suitable for estimating operating performance in ARP. If future ARP feed solutions have sorbate concentrations outside of the ranges reported in Tables 3-1 and 3-2, the expected MST performance should be reassessed.

Note also, that there are considerable differences among the three datasets in the range of sorbate concentrations. The ratio between the minimum and maximum sorbate concentrations vary from 1.67 for plutonium (lowest variation), 3.33 for uranium, 38.1 for strontium and 270 for neptunium (highest variation). Given this wide of range in sorbate concentrations, particularly for neptunium and strontium, DF values can differ considerably for solutions that have the same ionic strength (i.e., same sodium concentration), but different initial sorbate concentrations.

The TTR requested that SRNL provide DF values for 6 and 8-hour contact times. However, the available datasets have much more data at a 4-hour contact time and very little data at the 6-hour contact time. With concurrence from TF&CE, we evaluated DF values for each of the sorbates at 4-hour, 6-hour and 8-hour contact times and compared these values to that determined at 12 hours, which represents the current contact time in the ARP facility.

[Na] M	[Pu] ug/L	[Np] μ g/L	[U] μ g/L	$[Sr]$ ug/L	Reference
4.5	168	21,000	14,800	63.6	
4.5	63.5	406	9,020	90.1	12
5.4	114	457	9260	1,680	13
5.6	191	405	9,010	90.5	10
5.6	254	129	10,200	1,580	11
5.6	218	461	9,550	484	11
6.8	142	487	4,010	2,420	13
7.5	280	35,000	24,600	106	9

 Table 3-2. Initial Sorbate Concentrations in MST Performance Testing

Tables 3-3 through 3-6 provides a listing of measured and calculated DF values and single standard deviation at MST concentrations of 0.2 and 0.4 g/L. Calculated DF values were derived for selected contact times by fitting the DF versus contact time data to either a linear or second order polynomial function for each data set. These graphs and the mathematical equations are provided in the Appendix for each sorbate.

With the exception of uranium, the general trend is that the DF value decreases with a decrease in the MST concentration, a decrease in the contact time and an increase in the sodium concentration of the waste solution. For example, four of the datasets (Phase IV, Phase V and two set from Phase II mMST) measured MST performance at both 0.2 g/L and 0.4 g/L and, thus, provide the best comparison for evaluating the influence of MST concentration. Comparing the DF values at each of the contact times within each data set indicated that decreasing the MST concentration from 0.4 to 0.2 g/L decreased the DF value by a factor ranging from 1.6 to 2.6, with a majority of the results very close to a value of two. A decrease in the DF value by a factor of two would be expected assuming no other parameters (e.g., changes in other sorbate concentrations) are influencing plutonium removal.

For plutonium, shorter contact times resulted in reduced DF values for all but one of the data sets. For example, in the Phase IV⁶ and Phase V^4 datasets, the plutonium DF value at a contact time of 4 hours is about 60% of that measured or calculated to be at 12 hours. At a contact time of 8 hours, the DF value is about 82% of that at 12 hours for single contacts with either 0.2 or 0.4 g/L MST. In the Phase II mMST data sets with simulated and actual tank wastes, the impact of reduced contact times proved smaller at the 4-hour and 8-hour contact time DF values measured about 80% to 100% of the 12-hour values at both MST concentrations.

Comparison of the DF values at each of the contact times for the salt solutions having increasing sodium concentrations revealed that the plutonium DF value at 5.6 M Na is about a factor of two lower than that measured at 4.5 M Na. For a 7.5 M Na salt solution the plutonium DF value is decreased by a factor of about four compared to that measured in a 4.5 M Na salt solution.

Table 3-5. Average neptunium DF values upon contact with 0.2 and 0.4 g/L MST. Number in the bracket identifying the data set is the reference number from which the experimental data was obtained.

For the removal of uranium the measured DF values at the shorter contact times were not statistically different than those measured at 12-hours with either 0.2 or 0.4 g/L (see Table 3-4). Furthermore, the DF values were not influenced by Na ion concentrations ranging from 4.5 M to 7.5 M at contact times of between 4 and 12 hours. Note that even though the DF values for uranium are low compared to other sorbates, on a mass basis, uranium loading onto MST is much higher than the other data sets except for the Phase III datasets where the initial uranium and neptunium concentrations were both very high. Because of the relatively high mass concentration of uranium (typically 10,000 ug/L), measureable differences in uranium concentrations are not observable at contact times of 12 hours or less with MST concentrations of 0.2 and 0.4 g/L in salt solutions having Na concentrations ranging from 4.5 M to 7.5 M.

For neptunium, most of the DF values measured at 0.2 g/L MST were about a factor of 1.3 to 1.5 times lower than that measured at 0.4 g/L MST indicating less neptunium removal at the lower MST concentration. The results indicate that at each contact time, the percentage of neptunium removed from solution is about 10% less upon addition of 0.2 g/L MST compared to 0.4 g/L . This result suggests more effective removal of neptunium at the lower MST concentration, which is confirmed by comparing the calculated loading of neptunium onto to the MST. Neptunium loadings at 0.2 g/L MST measured between 1.08 and 2.12 times that at 0.4 g/L MST.

The rate of neptunium removal followed the same general trends observed with plutonium. At an MST concentration of 0.2 g/L, the DF values at a contact time of 4-hours and 8-hours reached on average 85% and 90%, respectively, of that measured at 12 hours. At an MST concentration of 0.4 g/L, the DF values at a contact time of 4-hours and 8-hours reached on average 69% and 83%, respectively, of that measured at 12 hours.

The very large difference in neptunium concentrations across all of the data sets makes it difficult to quantify the influence of ionic strength on the performance of MST to remove neptunium. Inspection of the DF values measured in the Phase III data set revealed no difference in the DF values from the 4.5 M Na and 7.5 M Na solutions. Note however, the initial neptunium concentrations in these two solutions varied considerably: 10,500 ug/L in the 4.5 M Na solution and 35,000 ug/L in the 7.5 M Na solution. The Phase IV and Phase V testing featured solutions having nearly identical neptunium concentrations, 406 ug/L and 403 ug/L, and Na concentrations of 4.5 M and 5.6 M, respectively. Comparison of the DF values in these data sets indicated a small decrease in the DF values at each contact time upon an increase in the Na concentration. It would be expected that the DF values would further decrease in salt solutions having higher Na concentrations due to the increased competition between Na^+ and the neptunyl ion, NpO_2^+ for sites on the MST.

Strontium removal in these data sets proceeded to no lower than 94.2% (DF = 17.1) removal and as high as 99.8% ($DF = 499$) removal within 12 hours of contact with MST indicating that MST has high affinity for strontium. In data sets with solutions containing less than 100 ug/L strontium $>99\%$ (DF >100) of the strontium was removed within 4 hours of contact with an MST concentrations of 0.2 g/L. Even with data sets having salt solutions with strontium concentrations of 1680 and 2,420 ug/L, strontium removal ranged from 94.2% (DF – 17.1) to 98.6% (DF = 69.1) with 0.2 g/L MST and contact times of 8 hours or less.

As with neptunium, the Phase IV and Phase V datasets featured solutions having nearly identical strontium concentrations, 90.1 ug/L and 90.5 ug/L, and Na concentrations of 4.5 M and 5.6 M. respectively. At each contact time and at both MST concentrations, the DF values decreased with an increase in the Na concentration. For the tests with 0.2 g/L MST, the difference was a factor of approximately 4.8. At an MST concentration of 0.4 g/L the difference in strontium DF values

was a factor approximately 3.6. Note, however, that even though the DF values varied by rather large factors, the change in total strontium removed was rather small, 94.2% versus 99.8%. It would be expected that the DF values would further decrease in salt solutions having higher Na concentrations due to the increased competition between Na^+ and Sr^{2+} for sites on the MST.

Table 3-7 provides a summary of the lowest DF value measured or calculated for each sorbate from each of the data sets reported in the preceding tables at MST concentrations of 0.2 g/L and 0.4 g/L. The standard deviation for the measured values, when available, is also reported. These values may be used to estimate the minimal fraction of each sorbate that would be removed from salt solutions ranging in Na concentrations from 4.5 M to 7.5 M and given the respective MST concentrations and contact times.

Table 3-7. Minimum DF Values and Single Standard Deviation for Plutonium, Uranium, Neptunium and Strontium in Alkaline Salt Solutions of Varying Ionic Strength Contacted with 0.2 g/L or 0.4 g/L MST

Average DF Value (Standard Deviation) with 0.4 g/L MST

3.2 Filtration Performance at Reduced MST Concentration

The proposed reduction in MST from 0.4 g/L to 0.2 g/L will affect the performance of the cross flow filter. Models have been used to predict cross flow filter performance as a function of operating parameters.14-16 One of the models, the Murkes and Carlsson model, is described by equation [1]

$$
J = \frac{B \Delta P}{\mu \left[\frac{c^2 \Delta P}{\left(av^3 + bv^3 \cdot 5 + 0.28v^4 \right)^{1/2}} + L_0 \right]}
$$
 [1]

where J is filter flux, B is Darcy's permeability constant for the filter cake, ΔP is transmembrane pressure, μ is liquid viscosity, v is axial velocity, L_0 is filter resistance, d_p is particle size, c_v is volumetric concentration of solids, D is filter tube diameter, ϵ filter cake porosity, $a = 21\mu/d_p\rho_I$, b $= 6(a/21)^{1/2}$, and $c = 2.43c_v^{1/3}[2D/(1-\epsilon)\rho_L)]^{1/2}$. The filter tube resistance, filter cake porosity, and filter tube diameter are assumed to be constant. Using a liquid viscosity of 2.5 cP, a particle size of 3 um, a liquid density of 1.26 g/mL, and axial velocities of 10 ft/s and 15 ft/s, the authors calculated the magnitude of the velocity terms in equation [1], and found the velocity cubed term to be the largest.

The authors simplified equation [1] and produced a new model described by equation [2]

$$
J = \frac{b_1 \Delta P v^{1.5}}{\left[b_2 \Delta P C^{2/3} d_p \rho_L^2 + \mu v^{1.5}\right]}
$$
 [2]

where J is filter flux (gpm/ft²), ΔP is transmembrane pressure (psi), v is axial velocity (ft/s), C is concentration (g/L), d_p is particle size, ρ_L is liquid density, and μ is liquid viscosity, and b_1 and b_2 are constants.

SRNL collected filtration data with simulated SRS waste and full-scale filter tubes (10 ft length, 5/8 inch ID, 0.5 um pore size) and fit the data with the model described by equation $[2]$.¹⁷ Since the same feed was used for each of the tests, the particle size, liquid density, and liquid viscosity are assumed to be constant, and equation [2] reduces to

$$
J = \frac{a_1 \Delta P v^{1.5}}{[a_2 \Delta P C^{2/3} + v^{1.5}]}
$$
 [3]

where a_1 and a_2 are constants.

According to equation [3] filter flux is a function of axial velocity, transmembrane pressure (TMP), and solids concentration. The impact of these factors varies based on the magnitude of each. At high axial velocity, low TMP, or low concentration, the second term in the denominator dominates, and the filter flux is proportional to TMP. At high TMP, high solids concentration, or low axial velocity, the first term in the denominator dominates, and filter flux is a function of axial velocity and solids concentration. At the ARP operating conditions $\left(\sim\right]$ 10 ft/s axial velocity, \sim 30 psi TMP, and 0.8 – 5 wt % insoluble solids), solids concentration has a significant impact on filter flux.

Figure 3-1 compares the model predictions with data from ARP. The model under-predicts the flux, but the effect of solids concentration is comparable between the model and the data. The model was fit with data from tests with a $0.5 \mu m$ filter. The ARP uses a $0.1 \mu m$ filter. SRNL testing with actual waste at bench scale and simulated waste at pilot-scale showed the 0.1 µm filter produced a larger flux than the 0.5 μ m filter.^{18,19} The increase ranged from 30% to 65%. Applying a correction factor of 1.4 to the model predictions shows better agreement with the ARP data. [Figure](#page-19-0) 3-2 compares the predicted flux, using the correction factor, to the ARP operating data. The agreement is good.

Figure 3-1. Comparing Filter Flux Prediction with ARP Data

Figure 3-2. Comparing Corrected Filter Flux Prediction with ARP Data

Models are usually fit to the specific system to which they apply. The lack of variation in the axial velocity and TMP at ARP does not lend itself to modeling their effects on filter flux with the ARP data.

By applying the model trend with a multiplier of 1.4, the reduction of MST concentration from 0.4 g/L to 0.2 g/L is expected to have the following impacts on production. Starting with lower solids will result in an increased filtration rate as predicted by the model. Decreasing from 0.4 g/L to 0.2 g/L MST is predicted to increase the initial average filtration rate by approximately 10%, or approximately 40 minutes per cycle on average relative to the current performance for the initial cycles. Each cycle will add one-half of the additional solids added per the current batches. Therefore, during operation, the concentration of solids that are fed to the filter will increase more slowly resulting in the filter operating longer at lower solids concentration and the corresponding higher flux.

Thus filter performance is expected to generally follow the prediction shown in [Figure](#page-19-0) 3-2 with filtration rate decreasing as insoluble solids concentration increases during concentration. The filtration rate will decrease and eventually follow the previous production curves but with more feed processed to advance on the concentration curve. The net result will be an initial higher volume of production, eventually matching current production rates. Since other parameters (TMP and axial velocity) are expected to remain consistent with previous operation, the slower increase in solids will result in longer durations at the filtration rates. In other words, the increase in concentration is driving the reduction in flux. Reducing the addition of MST per batch has the net effect of slowing the solids addition rate. That will result in a decrease in flux drop as a function of time, and a higher average filtration rate for each batch.

Although the model has not been fit to ARP operation, it can be used to estimate of potential impact to process time of reducing the solids loading to a crossflow filter. Using the model to predict the time required to process 50 batches of 3800 gallons of material for the current and reduced MST cases would result in a filtration time savings of up to 20%. This value was obtained by assuming that the only solids are MST; supernate, axial velocity (10.5 ft/sec) , and TMP (30 psi) are the same in both cases. Final solids would be an average of 5 wt % for the 0.4 g/L MST case and 2.5 wt % for the 0.2 g/L MST case. The model prediction of a complete cycle with 50 batches is approximately 460 hours of operational time $(0.4 \text{ g/L MST case})$. This is comparable to actual data (Salt Batch 3 cycle 2 required 448 hours to process \sim 190K gal in 50 batches).

TMP per the data is approximately 30 psi. Filter theory says that lower filtration rates will result in higher throughputs. At half the filtration rate, the pressure drop will be half as well (Note: at high enough pressure, flux becomes independent of pressure). The filter will operate longer prior to reaching the limiting pressure drop. If cleaning is based on a set schedule (number of batches produced) there may be an opportunity to optimize TMP to increase production rate by reaching a limiting pressure drop at the same time as planned cleaning. One caution is that increased TMP may lead to less efficient cleaning of the filter. This would need to be evaluated. In addition, the expected increase in flux at the initial lower solids may allow for a reduced TMP for those first batches. There may be an opportunity to reduce TMP at low solids though still resulting in a higher flux than historically seen at higher solids. By reducing TMP, the higher production rate could be sustained for a larger volume of product.

Inspection of the data provided from Salt Batches 3 and 4 shows a decrease in TMP during operation. It was concluded that this is due to increased resistance in the guard filter. Ideally, the guard filter and the primary filter should foul at approximately the same rate. This is typically done by adding additional filter area or loading capacity.

4.0 Conclusions

TF&CE is evaluating changes to the ARP operations to decrease the MST concentration to 0.2 g/L and contact time to between 6 and 8 hours. For this evaluation, SRNL reviewed previous datasets investigating the performance of MST at 0.2 g/L in salt solutions ranging from 4.5 to 7.5 M in sodium concentration. In general, reducing the MST concentration from 0.4 to 0.2 g/L and increasing the ionic strength from 4.5 to 7.5 M in sodium concentration will decrease the measured DF values for plutonium, neptunium, uranium and strontium. Conservatively, low DF values for each sorbate were obtained from a review of the available datasets at contact times of four and eight hours (see Table 3-3). These DF values are applicable within the sorbate and sodium concentrations used in the experimental measurements.

Decreasing the MST concentration in the ARP from 0.4 g/L to 0.2 g/L will produce an increase in the filter flux, and could lead to longer operating times between filter cleaning. The increase in flux is a function of a number of operating parameters, and is difficult to quantify. A general estimation has been made based on filtration theory of up to a 20% increase in filtration time as a result of decreasing MST solids.

5.0 Recommendations, Path Forward or Future Work

Evaluate if increased TMP leads to less efficient cleaning of the filter.

6.0 Acknowledgements

The authors thank T. B. Peters for calculating the sorbate concentrations in mass concentration units for ARP/MCU Macrobatches $1 - 5$.

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Appendix

Figure A-3. Neptunium DF values versus time from salt solutions of varying sodium concentration. MST concentration equals 0.2 g/L (upper) and 0.4 g/L (lower).

Distribution:

S. D. Fink, 773 - A K. M. Fox, 999-W B. J. Giddings, 786 -5A C. C. Herman, 999 - W S. L. Marra, 773 - A F. M. Pennebaker, 773 -42A D.T. Hobbs, 773 - A D. T. Herman, 735 -11A M. R. Poirier, 773 -42A C. Wilson, 773 - A Records Administration (EDWS)

E. J. Freed, 704 -56H D. J. Martin, 241 -152H K. L. Lang, 704 -27S R. V. Doniphan, 241 -156H M. W. Geeting, 241 -152H C. J. Scherman, 241 -197H M. H. Barnett, 766-H T. A. Le, 766 - H A. R. Shafer, 704 -27S C. K. Chiu, 704 -27S S. E. Campbell, 241 -152H S. P. Mcleskey, 241 -152H B. A. Gifford, 704 -56H R. M. Wolfenden, 704 -56 H K. H. Subramanian, 766 - H K. D. Harp, 766 - H R. K. Leugemors, 992 -5W W. B. Brasel, 992 -2W

P. R. Jackson, DOE-SR, 703-46A